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**Investigation on the interference of
glyconitrile in cyanide
determination according to ISO 14403**

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Abstract

Glyconitrile was identified as a positive interference during the cyanide determination following ISO 14403. Raising the pH during sample preparation leads to the decomposition of glyconitrile and to a false positive cyanide signal. The O/I Analytical CNSolution 3000 is a high throughput system with automatic sampling and easy handling, which severely suffers from the glyconitrile interference. A new method is set up und validated.

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I also have to thank my family for the unconditional support.

1. Introduction

Cyanide is a well-known **toxic compound** with severe effects on humans and the environment. A dose of **1mg per kg body weight** can already be lethal for humans, because the cyanide ion is directly inhibiting the cellular respiration (cytochrome oxidase).¹ It is frequently found in industrial as well as in municipal wastewaters, and therefore this compound is strictly regulated by public authorities. The limits for the drinking water content in the US and the EU are 0.2mg/L and 0.05mg/L, respectively.² Cyanide can occur in **various species**, which have different toxicities. This is primarily caused by the stability of the compounds that are formed from cyanide and the large spectrum of other substances. Two cyanide-containing fractions are distinguished in literature:

WAD cyanide – Weak Acid Dissociable cyanide²

This group of cyanides contains the most poisonous species. It consists of the **free cyanides**, which are hydrocyanic acid (HCN) and the cyanide ion (CN⁻). Further, the **weak cyano-complexes** of zinc, copper, cadmium, mercury, nickel and silver belong to this class³. All these compounds have in common, that they can release cyanide at **pH 4**. Other frequently used terms with the same meaning are **available** and **easily released cyanide**.

Total cyanide^{2,4}

Some cyanide containing compounds are rather stable and therefore have a lower toxicity. This especially applies to **organic cyanides** and **strong cyano-complexes** (of e.g. iron, cobalt, gold and platinum). The total cyanide is the sum of these and the WAD cyanides.

Because the fraction of the WAD cyanides has the highest impact on the environment, its determination is very important. Many decisions regarding the process management and wastewater treatment do directly rely on accurate analysis results. The aim of this work is to study interferences on methods used for the determination of WAD cyanides in order to ensure, that the analysis can be done exactly and without false detection of cyanides that are not part of the WAD fraction.

2. Methods

Due to the importance of WAD cyanides several regulative bodies provide procedures for the determination of this parameter. Two norms will be discussed in this chapter.

2.1. Description of ÖNORM M 6285⁴

Aim/general

This method contains procedures for the determination of **WAD-** and **total cyanide** from urban and industrial wastewaters. This work took a closer look on the analysis of free cyanide. It is performed by **acidic release** of HCN from the matrix, **absorption in alkaline solution** and **photometric quantification** by the pyridine/barbituric acid method. This method is regarded as the **reference technique** for the determination of WAD cyanide. This method is suitable for samples containing between 0.02mg/L and 0.25mg/L WAD cyanide.

Apparatus and procedure

If samples are **not analyzed immediately** after collection, 5mL sodium hydroxide (5M), 10mL phenolphthalein (0.03g phenolphthalein in 90mL ethanol and 10mL trichloromethane) and 5mL tin(II)chloride solution (50g SnCl₂*2H₂O dissolved in 40mL 1M HCl; filled up with DI water to 100mL) have to be added per liter. The pH has to be adjusted to 8 with 1M NaOH solution, before 10 mL zinc/cadmium sulfate solution (100g ZnSO₄*7H₂O, 100g 3CdSO₄*8H₂O in 1000mL DI water) are added. Samples should be kept in a cool and dark place.

For the **release** and **absorption** of hydrocyanic acid, 10mL zinc/cadmium sulfate (see above), 10mL EDTA (100g ethylenediaminetetraacetic acid dissolved in 940mL DI water) and 50 mL buffer solution (80g potassium hydrogen phthalate dissolved in 920mL DI water), 0.3g zinc dust (>98% of particles <62µm) and 100mL of fresh or stabilized wastewater have to be added to a round bottom flask. After adjusting the pH to 3.9±0.1 with 1M HCl or 1M NaOH, the apparatus has to be closed immediately. (This part was done differently in this work, because it is not practicable in every day routine analysis of multiple samples. Following the common practice in the industry, sample and reagents are mixed in a beaker. Also, the pH adjustment is done before the

transfer to the release and absorption apparatus.) The absorption vessel has to be filled with 10 mL of 1M NaOH solution before the air throughput is set to 30 to 60 L/h. After four hours, the extraction is complete.

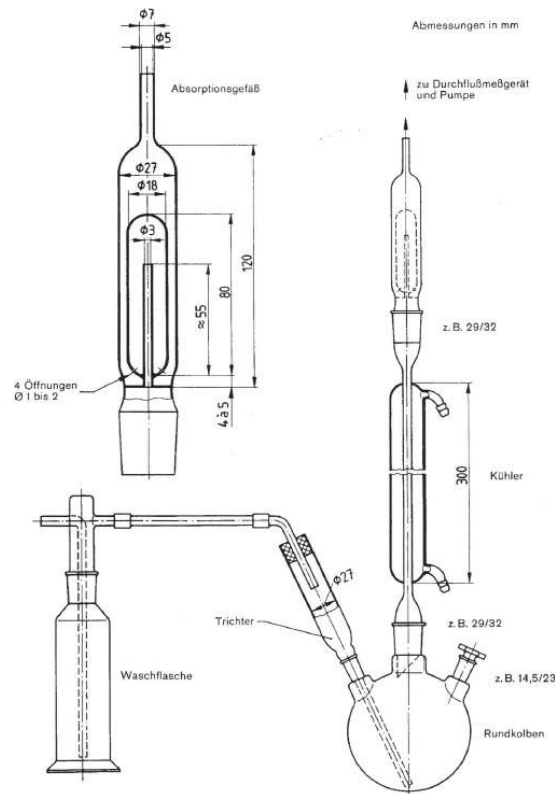


Figure 1: Apparatus for the release and absorption of WAD cyanide⁴

In the last step, the amount of absorbed cyanide is determined **photometrically** by the **pyridine/barbituric acid method**. Therefore the 10mL absorption solution has to be transferred to a 25mL volumetric flask, which is then made up to volume. 10mL of this solution is pipetted to another 25mL volumetric flask. Then 2mL buffer (pH 5,4; 5g NaOH; 11.8g succinic acid filled up to 100mL with DI water), 4mL HCl (1M) and 1mL chloramine-T solution (N-chloro 4-methylbenzenesulfonamide sodium trihydrate filled up to 50mL with DI water) are added. The flask is then closed and left for five minutes. 3mL of the pyridine/barbituric acid reagent (3g barbituric acid, 15mL pyridine, 3mL HCl 1.12 g/mL will up to 50mL with DI water) have to be added to start the colorimetric reaction. The measurement should be done after 20 ± 5 minutes in 10mm cuvettes at 580nm.

Analytical performance

According to ÖNORM M 6258, the coefficient of variation found in a round robin test was **28%** for stabilized samples. During this work, **poor reproducibility** over the whole concentration range could be observed too. One reason may be the complex procedure, which provides many possibilities for the loss of analyte. As described later, especially the pH adjustment to low values can cause significant problems. In the literature, some of these problems were also described for similar methods.⁵

Advantages and disadvantages

This method can be done with **ordinary laboratory equipment** at relative **low cost**. However, Operators need to be trained before carrying out this procedure for the first time to minimize errors.

A significant disadvantage is the **long duration** of the **analysis** of at least 5 hours. This results in a very long response time, which is not desirable in wastewater management. The workload can also **not exceed 10 samples** per shift and operator, because all measurements have to start at the same time and need an individual apparatus.

2.2. Description of EN ISO 14403⁶ (OIA-1677³) and OI Analytical CNSolution 3000

Aim/general

The CNSolution 3000 (OI Analytical, United States) is a commercial instrument performing analysis according to the **USEPA method OIA-1677**. This method is consistent to EN ISO 14403, which regulates the characteristics of flow injection analysis with gas diffusion and amperometric detection for the determination of WAD cyanide.

It is a high throughput system with automatic sampling and easy handling. It shows **sufficient reproducibility** (see section 18.3 of OIA-1677³) and has a **wide operational range** (0.005 to 0.2 mg/L).

EN ISO 14403⁶

EN ISO 14403 regulates the determination of total and free cyanide using continuous flow analysis (CFA). A wide variety of samples is possible with a typical limit of detection of 3µg/L in an operational range of 10µg/L up to 100µg/L. Known interferences are oxidants (e.g. chlorine), sulfides, aldehydes and thiocyanate.

For the determination of total cyanide, a UV lamp is necessary to degrade complex bound cyanide. All following steps are the same for both total and free cyanides and are described below.

Part of this standard is also the regulation of CFA using continuous distillation.

CNSolution 3000 and OIA-1677³

This method does not need any complex preparation steps. The pH of samples has to be raised to 11.0±0.1 with 1M sodium hydroxide solution and solid components of the samples have to be removed by filtration. To ensure a stable instrument detector baseline, it is necessary to switch on the pump at least 30 minutes before the measurement starts.

Figure 2 shows the flow diagram of the OI Analytical CNSolution 3000. It is a modular system with four main components located in one housing. In addition, an OI Analytical 120-position autosampler is used. The PC based software “Winflow 4.0” is used to control the instrument and for the interpretation of the recorded data.

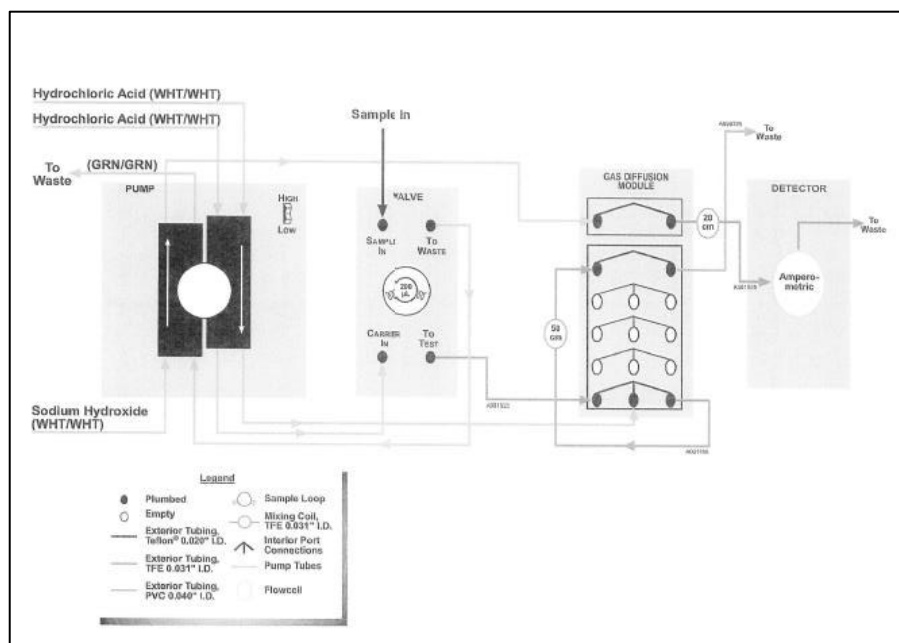


Figure 2: flow diagram of the CNSolution 3000 device for the measurement of WAD cyanide³

The **peristaltic pump** (Figure 3) has four channels. One transports 0.1M sodiumhydroxide solution, two contain 0.1M hydrochloric acid and the last one is connected to the “to waste” exit of the **six-port-valve** (Figure 4). This part is the second module in Figure 2 and does the sample switching. It has two inlets, one for the incoming stream from the autosampler and the other for the 0.1M



Figure 3: Peristaltic pump

HCl carrier stream. Two of the ports are connected to a 100 μ L sample loop and the last two ports are used as an exit to the waste container and the further analysis system.



The **six-port-valve** consists of two plates. One contains six ports that are connected in pairs through cannels located on the other plate. This results in two operational states, which can be selected by twisting of the plats against each other. During “loading”, the sample coming from the autosampler is flushed through the sample loop to the waste container. If the valve is switched to “inject”, the carrier stream will transport the sample from the sample loop into the “to test” outlet.

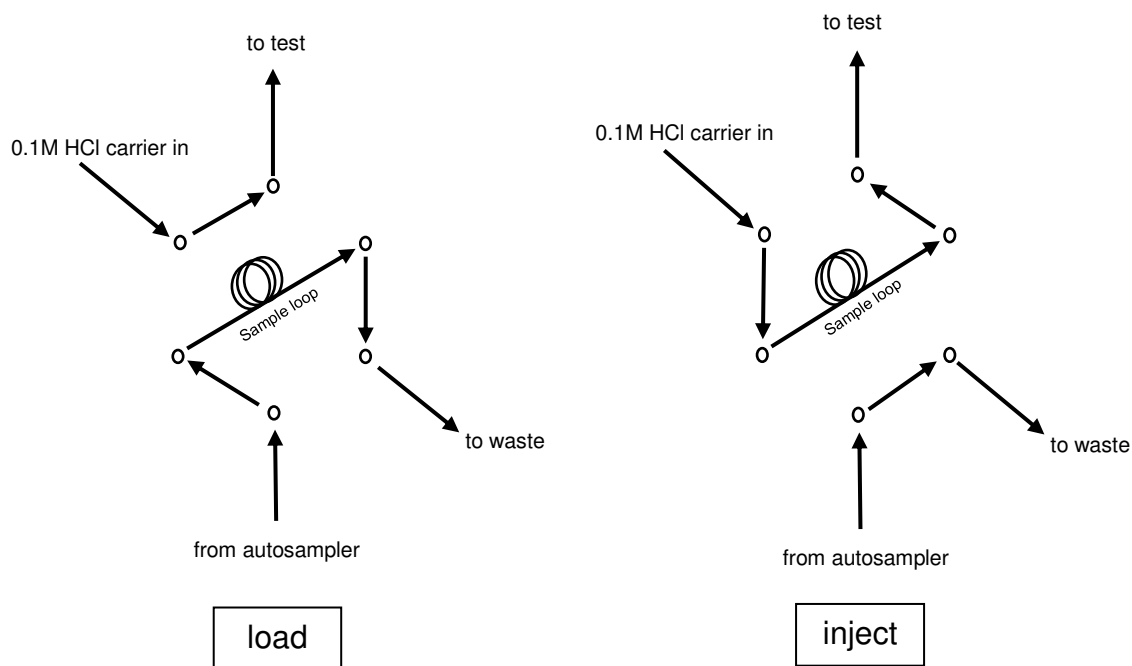


Figure 5: Operation states of a six-way-valve

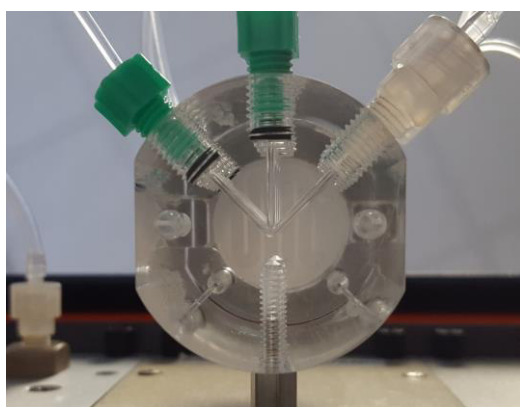


Figure 6: Mixing chamber

100 μ L sample are now transported by the continuous flow system to the **mixing chamber** (Figure 6) where it is acidified by the second 0.1M HCl stream. Investigations during this work showed, that the flow rate of the sample and the HCl are similar before mixing and that the resulting pH is close to 1 for samples up to pH 12.

The formed hydrocyanic acid is now transported to the **gas diffusion chamber** (Figure 7), where it is permeating through a hydrophobic polypropylen membrane into the 0.1M NaOH stream. In the alkaline medium, the cyanide ion is formed again. The amperometrical detection takes place in the **detector module** (Figure 8). It is a three-electrode assembly consisting of a silver working electrode, a

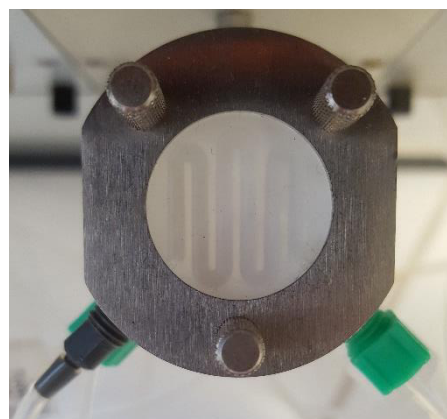


Figure 7: Gas diffusion chamber, top down perspective

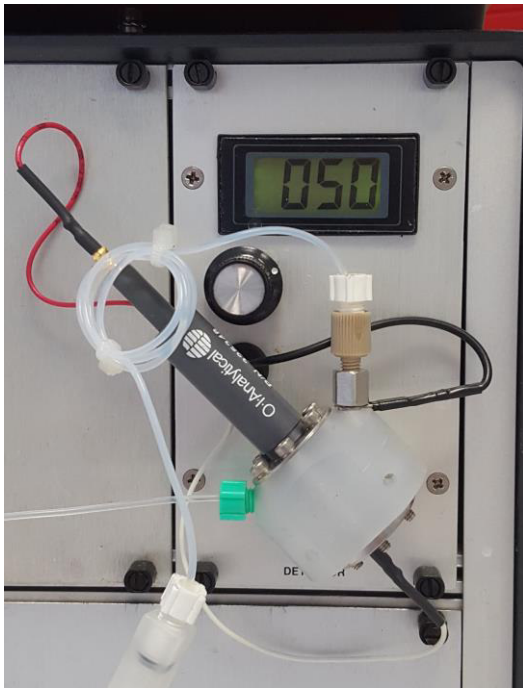


Figure 8: Detector module

silver/silver chloride reference electrode and a flow through stainless steel counter electrode. The applied potential is 0.050V.

The peak current is used for the calculation of the concentration, which is done automatically by the software “Winflow 4.0” using a two-point calibration.

In Figure 9 all of these parts/modules can be seen together. The above described mixing and the diffusion chamber are both located on the same module (third from the left). Additional pictures can be found in the appendix (section

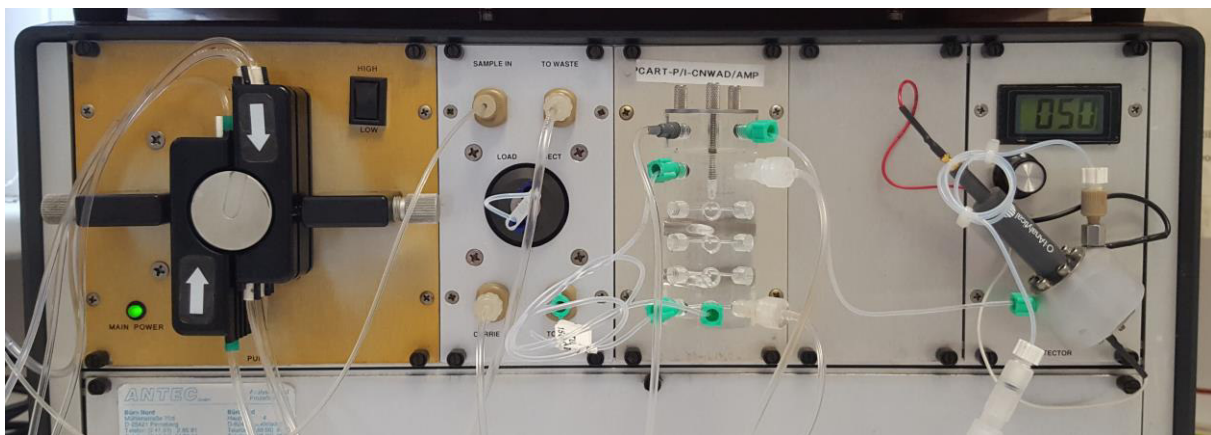


Figure 9: Complete setup of the OI Analytical CNSolution 3000

Analytical performance

Nine laboratories participated in a validation study to show the reliability of this method³. Various typical matrices were tested, but only the results for DI water with 0,01M NaOH (pH 12) should be mentioned here.

Table 1: Results of an interlaboratory validation study showing the reliability of OIA-1677

Sample	CN ⁻ Concentration [mg/L]	Average Recovery [%]	Rel. Standard Deviation [%]
DI water 0,01M; NaOH (pH 12)	0.1	108	4.0
DI water 0,01M; NaOH (pH 12)	0.2	101	8.0
DI water 0,01M; NaOH (pH 12)	10	103	2.0

During this work, the CNSolution 3000 showed lower standard deviations in most cases (for examples see appendix section 11.2.). In the comparison of different methods, a **relative standard deviation of 2%** was seen as a representative value for the calibration at 0.2mg/l CN⁻.

Advantages and disadvantages

A very clear advantage of this method is the **very short analysis duration**. The calibration of the instrument takes approximately 12min and the triple determination of one sample can be done in about 6min. The total time necessary for a set of five samples is less than 45min. Along with the necessary baseline stabilization before the measurement and the sample preparation, the total time between the unexpected arrival of a sample at the lab and the complete results is less than one and a half hour. In comparison to the ÖNORM M 6258⁴, the CNSolution 3000 has a **more complex setup**, which may fail at some point. In that case, troubleshooting should not be any problem, because of the easily accessible components.

Additional Information

During this work, the pH was not always set to 11 during the sample preparation. If the standard method was changed, it was always noted in the experiment description.

2.3. Comparison

Additionally to the already mentioned aspects, the **analytical performance** and the **measurement duration** are clear advantages of the CNSolution 3000 following OIA-1677 over the ÖNORM M 6258. In everyday use, it is a very practical system, which is capable of handling a **large variety of wastewater samples** and has a **wider operating range**.

Nevertheless, every laboratory using this system should also be capable of performing analysis according to the ÖNORM method. It is an easy and reasonable way **to confirm results** and **bridge device malfunctions**.

3. Problem description

During daily routine analysis of waste water samples, the CNSolution 3000 following OIA-1677 showed **systematically higher** results than the reference method ÖNORM M 6258⁴. The measured values were nearly **twice as high** and this behavior was observed over a long period. Device malfunction or an analytical error could be excluded by careful investigation of each step. It was also noticed, that **increasing the pH value** during sample preparation **raised the difference in the obtained values** between the two methods.

Because of this issue, the legal emission limit of 0.1mg/L WAD cyanide can be easily exceeded while the actual concentration in wastewater is much lower. Therefore, the accurate quantification of the WAD cyanide concentration is crucial, because it is directly linked to decisions regarding process management and wastewater treatment.

4. First investigations with known interferences

A set of initial experiments was performed to gain insights into the factors that might affect the differences between the two norm methods for determining WAD cyanide. Some of the experiments were already conducted during the setup of the OIA-1677³ method, but they were redone in this work to ensure reproducible and consistent results.

The following list is not in a chronological order. They were done at different times and do not refer to each other.

4.1. Dissociation of Non-WAD cyanide complexes

The aim of this experiment was to quantify the amount of cyanide released from strong cyanide complexes of metals such as iron or cobalt.

Both cyanide complexes of **iron** were investigated during this work. They could not be synthesized in situ and therefore have to be dissolved from solid state (p.a. grade). Investigations were done with and without the presence of a known amount of free cyanide. The amount of added **hexacyanoferrate (II)** and **(III)** is given in mg/L of CN bound in the complex. This concentration would only be encountered, if a total cyanide analysis would have been performed. The investigated concentration range was rather wide, because the effect of these complexes is only noticeable at high levels. Following tables contain the measurement results for potassium hexacyanoferrat (II) and (III) from triple determinations.

*Table 2: Interference study with potassium hexacyanoferrate (II);
Values for 100mg/L were eliminated because of high RSD; n=3; SD not shown for clarity;*

mg/L CN ⁻ from K ₄ [Fe(CN) ₆]	Signal [pA]	c(CN ⁻) [mg/L]	RSD [%]	Recovery [%]
0.5	1504	0.005	2.8	1
1	3061	0.009	2.1	0.9
2	5797	0.018	2.1	0.9
25	60825	0.188	1.4	0.8
100	32594	0.101	4.9	0.1

The results listed in Table 2 show, that the effect of strongly-bound cyanide on the determination of WAD cyanide is negligible. Less than 1% of cyanide added as strongly-bound cyanide was detected by the used method. Therefore, potassium hexacyanoferrate (II) cannot cause positive interference. The result of the 100 mg/L sample was not taken into account, because the very high concentration may distort the measured value.

Table 3: Interference study with potassium hexacyanoferrate (II) in the presence of free cyanide; n=3; SD not shown for clarity;

mg/L CN ⁻ from K ₄ [Fe(CN) ₆] + 0.1 mg/l CN ⁻	Signal [pA]	c(CN ⁻) [mg/L]	RSD [%]
0.5	28377	0.088	0.5
1	20898	0.065	0.9
2	34250	0.106	2.2
25	81826	0.253	2.2

In the presence of free cyanide, it was also noticeable that only high concentrations of hexacyanoferrate (II) resulted in significantly higher cyanide values. However, at low hexacyanoferrate (II) concentrations unexpectedly low spike recoveries of only 65 % were encountered. The reason for this is unknown and further experiments would be needed to clarify this point.

Table 4: Interference study with potassium hexacyanoferrate (III) ; n=3; SD not shown for clarity;

mg/L CN ⁻ from K ₃ [Fe(CN) ₆]	Signal [pA]	c(CN ⁻) [mg/L]	RSD [%]	Recovery [%]
0,5	974	0.003	3.6	0.6
1	1949	0.006	3.5	0.6
2	3476	0.010	0.8	0.5
25	28453	0.086	2.5	0.3
40	16748	0.052	n.a.	0.1
60	23626	0.073	n.a.	0.1
80	32040	0.099	n.a.	0.1
100	38792	0.117	4	0.1

Table 5: Interference study with potassium hexacyanoferrate (III) in the presence of free cyanide; n=3; SD not shown for clarity;

mg/L CN ⁻ from K ₃ [Fe(CN) ₆] + 0.1 mg/L CN ⁻	Signal [pA]	c(CN ⁻) [mg/L]	RSD [%]
0,5	35027	0.106	0.2
1	34891	0.105	0.6
2	37544	0.113	0.8
25	68136	0.205	4.3

Potassium hexacyanoferrate (III) did not cause any interferences at concentrations typically encountered in waste water samples. The cyanide recoveries always stayed below 1%.

Cobalt forms another strong cyano-complex, which was investigated. For this test, solutions of cobalt and cyanide are mixed one hour before the analysis. The used amount of Cobalt is measured in molar equivalents in regard to the initial concentration of cyanide (0.2 mg/L).

Table 6: Interference study with cobalt

Equivalents of Co ²⁺	Signal [pA]	c(CN ⁻) [mg/L]	Recovery [%]
+0.5 eq.	8.190	0.024	12
+1 eq.	5.553	0.016	8
+2 eq.	4.701	0.014	7
+2 eq.	5.772	0.017	8.5

The result of this experiment was not clear. Either the cyano-complex was not formed completely or a portion of the cobalt complex is decomposed and the released cyanide could be detected. Either way, the recovery rates were around 10%.

Conclusion:

Hexacyanoferrates were found to cause a positive interference on the WAD cyanides at very high concentrations. However, typical samples do not contain 100 to 1000 times more total than WAD cyanide. Therefore, it is **very unlikely** that these complexes caused the difference between ÖNORM 6285 and OIA-1677. This also holds true for cyanocomplexes of cobalt.

4.2. Thiocyanate and Cyanate

The potential cyanide release of **thiocyanate** and **cyanate** was also investigated. Even though it was considered unlikely, that these compounds caused the observed differences between the two norm methods.

Synthetic standards are produced by adding solid KSCN and KCNO to DI water. The concentrations were calculated in mg/L CN⁻, which theoretically could be released.

The following tables show the results of this experiment:

Table 7: Interference study: Thiocyanate; n=3; SD not shown for clarity;

mg/L CN ⁻ from KSCN	Signal [pA]	c(CN) [mg/L]
0.5	65	<LOD
1	266	<LOD
2	365	<LOD
25	500	<LOD
100	818	0.003

Table 8: Interference study: Cyanate; n=3;

mg/L CN ⁻ from KCNO	Signal [pA]	c(CN) [mg/L]
0.5	-129	<LOD
1	-55	<LOD
2	104	<LOD
25	188	<LOD
100	187	<LOD

Conclusion:

Thiocyanate and isocyanate **did not interfere** with the determination of cyanide using the CNSolution 3000.

4.3. Flocculation agent⁷

Some wastewater treatment procedures contain the application of **flocculation agents** such as **VTA EA 83**. It is an anionic polyacrylamide compound, which is used to accelerate the solid – liquid separation during sedimentation and flotation. It was also suspected to influence the cyanide determination by either releasing cyanide ion or interfering with the electrochemical detector (similar to sulphide, see section 4.5)

Two sets of standards were measured during this experiment. In the first set of experiments the amount of flocculation agent remained constant, while the cyanide concentration was varied. In the second set, the cyanide concentration remained constant, while the flocculation agent was varied.

The following tables show the results in detail.

Table 9: Interference study flocculation agent (floc.a.), cyanide concentration variation; n=3; SD not shown for clarity;

Target c(CN ⁻) [mg/L]	c(floc. a.) [mg/L]	Signal [pA]	Measured c(CN) [mg/L]	Difference [%]
0.01	1.5	4323	0.012	23
0.02	1.5	8021	0.023	14
0.05	1.5	17683	0.050	1
0.10	1.5	33416	0.095	-5
0.15	1.5	49458	0.141	-6
0.20	1.5	65800	0.188	-6
2.00	1.5	646058	1.843	-8

Table 10: Interference study flocculation agent, constant cyanide concentration; n=3; SD not shown for clarity;

Target c(CN ⁻) [mg/L]	c(floc. a.) [mg/L]	Signal [pA]	Measured c(CN) [mg/L]	Difference [%]
--------------------------------------	-----------------------	----------------	-----------------------------	-------------------

0.2	0.5	68242	0.188	-6
0.2	1	67541	0.186	-7
0.2	1.5	66275	0.183	-9
0.2	2	67358	0.185	-7
0.2	5	67247	0.185	-7

Conclusion: The presence of flocculation agent **did not increase** the measured cyanide concentration. However, spike recoveries in the range of 91 – 94% indicate a potential influence of the flocculation agent on the electrochemical cyanide determination. These low results might also be caused by the production procedure of the samples and the instrument inaccuracy.

4.4. Acetonitrile

Acetonitrile was also investigated during this work, because it is a potential component of wastewater, too.

It was tested in multiple concentrations at pH 8.3 and 11 with the CNSolution 3000. One sample was also analyzed according to the ÖNORM M 6258 at a release pH of 3.2 and 3.9.

The Acetonitrile concentration in the following table refers to the maximum releasable amount of cyanide. For example, 0.01 mg/L cyanide can be released from a standard containing 0.014 mg/L acetonitrile.

Table 11: Interference study acetonitrile; n=3

Acetonitrile [in mg/L contained CN ⁻]	CNSolution 3000		ÖNORM M 6258	
	~ pH 8.3 [mg/L CN ⁻]	pH 11 [mg/L CN ⁻]	~ pH 3.2 [mg/L CN ⁻]	pH 3.9 [mg/L CN ⁻]
0.01	<LOD	<LOD		
0.02	<LOD	<LOD		
0.05	<LOD	<LOD		
0.1	<LOD	<LOD		
0.15	<LOD	<LOD	<LOD	<LOD
0.2	<LOD	<LOD		

Conclusion:

In the investigated concentration range, **no effect of acetonitrile on the determination of WAD cyanide** was observed.

4.5. Sulfide

According to OIA-1677³ and literature on comparable techniques⁵, sulfide is the **most frequently mentioned positive interference**. Like cyanide, it is able to pass a hydrophobic membrane from the acidic to the alkaline side (H₂S is hydrophobic) and it leads to a similar signal at the detector as cyanide.

Following OIA-1677, water samples containing sulfide should be treated with lead carbonate to form insoluble lead sulfide. In this form, sulfur does not cause any more problems if the precipitated PbS and the excess of PbCO₃ is filtered off immediately (see below).

According to experienced lab staff, this method works well in most cases. Sometimes however, the positive interference could not be eliminated. This phenomenon should be investigated via the following experiment.

Lead carbonate treatment of wastewater samples

Cyanide in wastewater samples was quantified with the CNSolution 3000 at pH values from 10.5 to 12.5 with and without the addition of PbCO₃. Additionally, the results were compared to values from the ÖNORM method and the separately performed routine

method (also done with CNSolution 3000). Additionally the sulfide concentration was determined spectrophotometrically.

The aim of this experiment was to investigate if the removal of sulfide has any influence on the pH dependence of the measured signal.

Procedure:

Five beakers were filled with approximately 80ml of the same sample. After the pH value was adjusted to 10.0, 10.5, 11.0, 11.5 and 12.0 with 1M NaOH, the first measurement with the CNSolution 3000 was conducted. To remove sulfide from the samples, a small amount (one spatula tip; ~0.1g) of PbCO_3 was added to the samples. This addition needs to be done immediately before the measurement, because thiocyanate might be formed from the precipitated lead sulfide and cyanide. During this experiment, the stirring time was about 1 minute.

Evaluation:

For two weeks, samples from three different origins were analyzed daily in four different ways (CNSolution 3000, CNSolution 3000 daily routine analysis, "Skalar" continuous measurement device, ÖNORM M 6285). A large set of data was generated, but only limited knowledge could be extracted from it. Some measurements were discarded, because of failure during the analysis or statistical issues (e.g. high coefficient of variation). The remaining results had no uniform behavior. Some of them showed the same issues as observed during the routine analysis (see section 4.6). However, others represented the complete opposite.

Conclusion:

This investigation indicates a strong matrix dependent behavior of the PbCO_3 treatment that was not always able to eliminate the positive interference from sulfides or other matrix constituents. It was further considered likely, that **another component** in the matrix, also interfered with the cyanide determination leading to higher signals. The influence of sulfide on the cyanide measurement needed in any case a more detailed investigation.

4.6. pH dependence of the positive interference

During earlier investigations on the positive interference of the sample matrix on the cyanide signal, a pH dependence of the measured signal was observed. The concentration increased with the pH value, which was adjusted during sample preparation. This was especially strange, because the sample is acidified with 0.1M HCl immediately after injection to the system. Therefore, the difference in pH in the diffusion chamber can be expected to be very low. There was no plausible explanation how any of the known positive interferences (e.g. sulfide) can cause this behavior.

The following diagram shows an example for this pH dependence, done with spiked, real wastewater samples. It was measured during the sulfur investigations (see 4.5). The results are normalized to pH 11 to show the deviation from the measurement following ISO 14403.

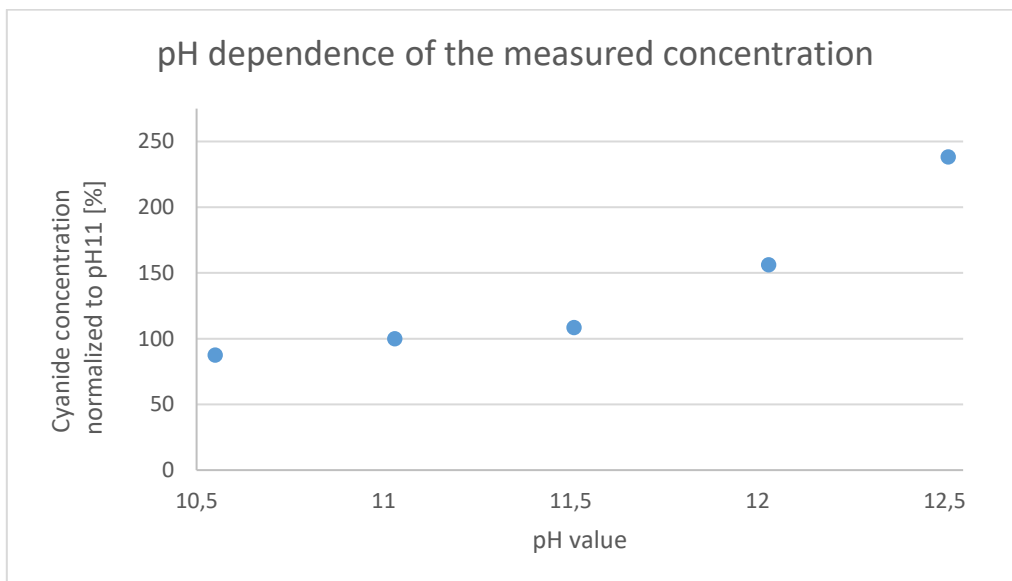


Figure 10: pH dependence of the measured signal (CNSolution 3000);
RSD in all cases smaller than 2%

5. The Cyanohydrin-hypothesis

5.1. Introduction

Because some possible interferences had been eliminated (see chapter 4), a more general literature study was started. In the course of this, the **Degussa-treatment**⁸ helped developing a new theory. It is a method developed to eliminate cyanide from washing waters of waste gas scrubbers.

The Degussa treatment is a staged procedure: At first, dust and other coarse particles are removed from the waste gas by cyclones and electro filters. Then, fine cleaning is done with gas scrubbers. The resulting wastewater contains cyanide or metal-cyanide-complexes, which are treated with formaldehyde at pH of about 7 (slightly acidic to alkaline). The reaction product is **glyconitrile HOCH₂CN**. In the next stage, it is hydrolyzed using H₂O₂ to obtain glycolic acid, which is easily **biodegradable**. Thereby, cyanide is removed from the wastewater.

Formaldehyde is a product of incomplete combustion, it could therefore also be present in gas scrubbing wastewaters. If the wastewater reaches **slightly alkaline** pH values, the **formation of glyconitrile** can be expected to take place to a certain extend as well. The concentration of cyanide is therefore lowered and glyconitrile potentially behaves different during the analysis according to ÖNROM 6285⁴ and EN ISO 14403⁶.

5.2. Literature research

This chapter presents the current status of **formaldehyde-interference** in the cyanide determination in literature. At first, a closer look is taken at ÖNORM M 6285, OIA 1677 and EN ISO 14403. Then a paper, which is focused especially on glyconitrile, is analyzed. In the last section, additional knowledge on the mechanism of glyconitrile formation is gathered and evaluated.

The consequences of all this information will be combined into the **Cyanohydrin-Hypothesis**, which is explained in section 5.3.

Glyconitrile (Figure 11⁹) is a compound formed by nucleophilic addition of formaldehyde and cyanide with the formula HOCH₂CN. It belongs to the substance class of

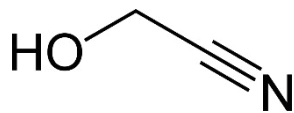


Figure 11: Structure of Glyconitrile (cyanohydrin of formaldehyde)⁹

Cyanohydrines (Figure 12¹⁰), which are obtained by the reaction of any aldehyde or ketone with cyanide. The functional region of the molecule

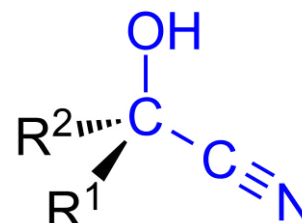


Figure 12: Functional group of Cyanohydrins¹⁰

consists of a **hydroxyl**- and a **cyano**-group, which are located on the same carbon atom.

5.2.1.A closer look at ÖNOR M 6285⁴, EN ISO 14403⁶ and OIA-1677³

After cyanohydrins and in particular glyconitrile had become the target of investigations, the methods ÖNOR M 6285, EN ISO 14403 and OIA-1677 were carefully reviewed.

ÖNROM 6285

In section 1 “Aim and area of application”, it is mentioned, that the presence of aldehydes and in particular formaldehyde can cause lowered measurement values.

In the course of the definition of different cyanides, (see 2. “Definitions”), it is indicated, that cyanohydrins are only partly determined as total cyanide and that nitriles in general are not part of the easily released cyanides.

However, it is also stated, that the behavior of cyanohydrin during the chemical fusion for the determination of total cyanide is not well understood.

OIA-1677

Aldehydes are listed as interferences (see section 8.5). The addition of ethylenediamine is suggested as a treatment method, but it is not described how this affects the measurement results. Further research on the mechanism is necessary.

EN ISO 14403

This method states, that organic cyanides should not be determined as free cyanide by this method.

Conclusion:

It is likely that aldehydes lower the cyanide concentration due to the formation of cyanohydrin. This compound seems to be instable at certain conditions (determination of total cyanide) and is measured as cyanide in this case.

5.2.2. Main literature

During an extensive literature search, only one paper covering matrix-induced interferences in the determination of cyanide had been found.¹¹ Unfortunately, the full text is in Japanese. Only the abstract and the labels of the figures are written in English. Nevertheless, it was very enlightening, because it linked the observed deviation between ÖNORM M 6285⁴ and OIA 1677³ with the presence of formaldehyde.

In this investigation, total cyanide is determined according to JIS K 0102¹², which is quite similar to the ÖNORM M 6285: Hydrogen cyanide is also released from the matrix and absorbed by a sodium hydroxide solution. The quantification is done spectrophotometrically with 4-pyridinecarboxylic acid-pyrazolone.

The authors believe that formaldehyde and cyanide contained in the samples react to form cyanohydrin. This results in lowered recoveries of total cyanide and a method is presented to avoid this. At first, the pH is raised to 12, which leads to **decomposition** of cyanohydrin back to cyanide and formaldehyde. To ensure no new cyanohydrin is formed during the analysis (at low pH during CN release), formaldehyde is eliminated with tetrahydroborat.

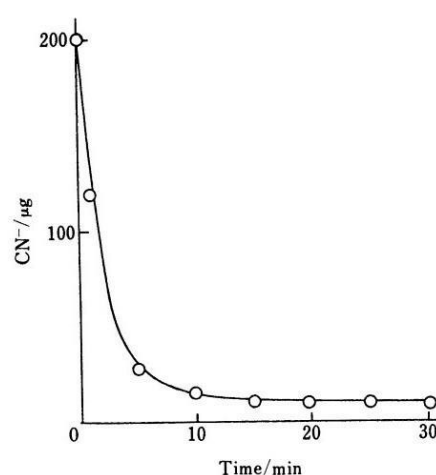


Figure 13: Time dependent formation of cyanohydrin at pH 7 (200µg CN⁻; 300µg HCHO)¹¹

The first figure in this paper depicts the **time dependence** of the cyanohydrin formation. The reaction is done at pH 7 with a slight molar excess of formaldehyde. This shows that the reaction needs at least **10 minutes** to reach equilibrium. This is rather quick but it is only reached a conversion factor of about **95 percent**.

Figure 14 shows the **equilibrium concentration** of the decomposition reaction at different pH values. It starts at 95% cyanohydrin at pH 7, which represents exactly the conditions of Figure 13. At pH 13, approximately 25% cyanohydrin remain in the sample. At pH11 (measurement pH of the CNSolution 3000) about half of the cyanide is released.

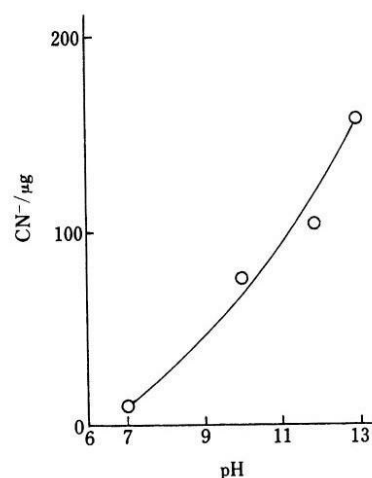


Figure 14: pH dependence of cyanohydrin (CN⁻:200 µg; HCNO: 300µg¹¹)

Figure 15 gives insight into the **pH-dependency** of the elimination reaction of formaldehyde with tetrahydroborane. At a pH of 12 and after a reaction time of 30 minutes, nearly all the cyanohydrin has been decomposed.

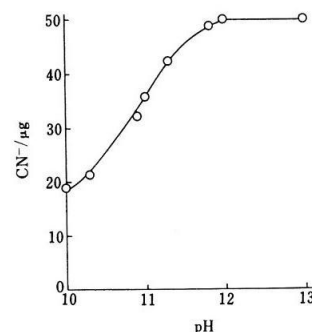


Figure 15: Effect of pH on the elimination of formaldehyde (CN⁻:50µg; HCHO: 500µg; NaBH₄: 0.3g;)¹¹

These three figures provide important information on the chemical behavior of the cyanide/formaldehyde/glyconitrile-system and are useful during the development of the **Cyanohydrin-Hypothesis** as explained later.

5.2.3. General Literature on Cyanohydrin

During the literature search, several publications treating the **formation reaction** of cyanohydrin were found. Most of them describe reactions for chemical synthesis, but their findings can be expected to also apply to wastewater matrices. The proposed mechanism is the **base catalyzed nucleophilic addition**.^{13,14,15} Polar solvents like water are especially beneficial, because the carbonyl group of the cyanohydrin is activated by interactions with OH.¹³ The equilibrium is not completely on the side of the products. The reaction is **reversible** and strongly **dependent on the pH**.^{13,14} Although the reaction is base catalyzed, **hydrogen ions stabilize** the formed alkoxide.^{13,16} Therefore a two-step mechanism is proposed: **1. Base catalysis 2. Acidic stabilization**.¹³

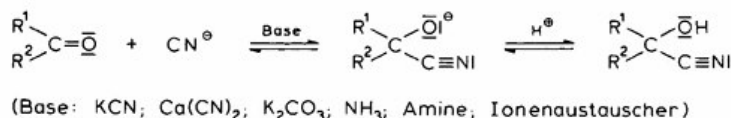


Figure 16: Reaction mechanism cyanohydrin formation¹³

The most stable products are formed by aldehydes, but the reaction will also take place with ketones.^{5,13,14,16,17}

Only one publication deals with **cyanohydrin as an interference** in the analysis of cyanide using the CNSolution 3000.¹⁷ Because in this case total cyanide is determined, cyanohydrins are seen as a negative interference. Ethylenediamin-treatment is the only suggested countermeasure, but it is claimed that this will only prevent additional formation of cyanohydrin. The recovery of cyanide from cyanohydrin is not possible.

Another issue observed during the cause of this master thesis is the rising cyanide concentration during storage. If samples are stabilized by raising the pH to 12 with sodium hydroxide, it seems reasonable to assume – based on the mechanism just discussed - that additional cyanide is formed.

5.3. Cyanohydrin-Hypothesis

In this section, all previously gained knowledge about cyanohydrins and in particular glyconitrile is merged into the **Cyanohydrin-Hypothesis**.

Cyanide and formaldehyde are both products of incomplete combustion and therefore both can be present in wastewaters side by side. These two species form cyanohydrin at a pH of 7 to 8 in an equilibrium reaction. The product can then be stabilized in a neutral or slightly acidic environment.

During the ÖNORM M 6258 analysis, the pH value is never raised above the original level of the sample. At a pH of 3.9 free cyanides are released to the gas phase, but glyconitrile remains in solution. In contrast, the sample preparation for the determination of available cyanides using the CNSolution 3000 includes adjusting the pH to 11 before the measurement. This leads to a shift of the reaction equilibrium to the side of cyanide and formaldehyde. As it is depicted in Figure 17, at pH values above the pKs of hydrocyanic acid the alcoholate of glyconitrile (pKs 16) is formed. This compound is less stable and therefore the reverse reaction is preferred.

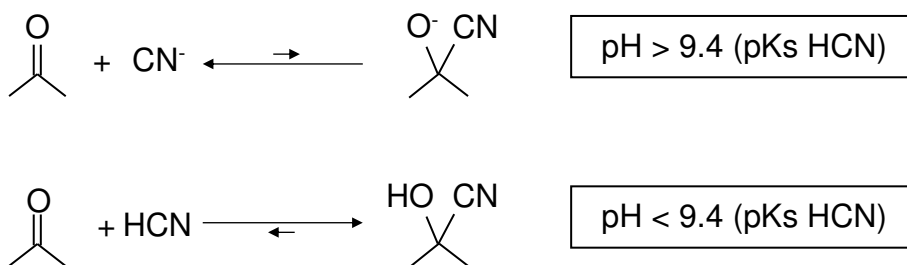


Figure 17: Formation reaction equilibrium of glyconitrile

At this point, it is important to note, that glyconitrile is not an interference in the classic sense. It is not a completely different substance like for example sulfide, which unintentionally leads to a signal at the detector. The decomposition of cyanohydrins causes a real increase of the cyanide concentration. However, cyanohydrin is per definition not part of the easily liberable cyanide fraction, because it cannot be released at a pH of 4². Further, it is only partly determined as total cyanide during ÖNORM M 6258 analysis.

Because of this, it is reasonable to consider cyanohydrins as interferences during the determination of WAD cyanides. Approaches to eliminate their influence on the measurement are therefore seen as legitimate.

6. Investigations on the Cyanohydrin-Hypothesis

To prove and extend the knowledge on the properties of cyanohydrin, a series of experiments with synthetic standards was performed. Glyconitrile in the form of a ~55% solution in water with ~0.5% phosphoric acid as stabilizer was used, purchased from Sigma Aldrich (~577g/L; product #: 374768¹⁸; lot #: BCBF4937V; CAS-number 107-16-4). This solution was originally intended to be used in synthesis applications, therefore its concentration was not exactly known. This had to be accepted, because it was not possible to find a p.a. grade source for this compound. Further, a concentration measurement could not be performed, due to the lack of an appropriate analytical technique. The calculation of the dilution steps are based on 55-masspercentage glyconitrile. Lacking the knowledge of the exact concentration of glyconitrile is certainly unsatisfactory however, all experiments using this stock solutions will provide a general trend that is only biased with a constant factor.

6.1. pH-dependence of the formation reaction and the reverse reaction

It is known from the literature, that cyanohydrin is only stable in neutral to acidic solutions, which was also confirmed by the fact that the stock solutions were stabilized with acid. Examined important factor is certainly how alkaline pH values influences the equilibrium between glyconitrile and cyanide.

Procedure

Six standards with a concentration of ~0.44 mg/L glyconitrile were produced by a three-step dilution from the 55-masspercentage stock solution. In the case of complete degradation of glyconitrile, this would have resulted in a cyanide concentration of 0.2

mg/L. Before the measurement, the pH values of the standards were adjusted to 7, 8, 9, 10, 11 and 12 using sodium hydroxide. Then they were measured with the CNSolution 3000 as quickly as possible (<2min). The calibration was done at pH 12, following the standard measurement procedure.

Results and Interpretation

Glyconitrile showed the behavior expected from the literature. Up to the pKs of hydrocyanic acid (9.4), the measured cyanide concentration remained at a rather low level. At higher pH, the degradation of glyconitrile increased rapidly and reached 80% at pH 11.

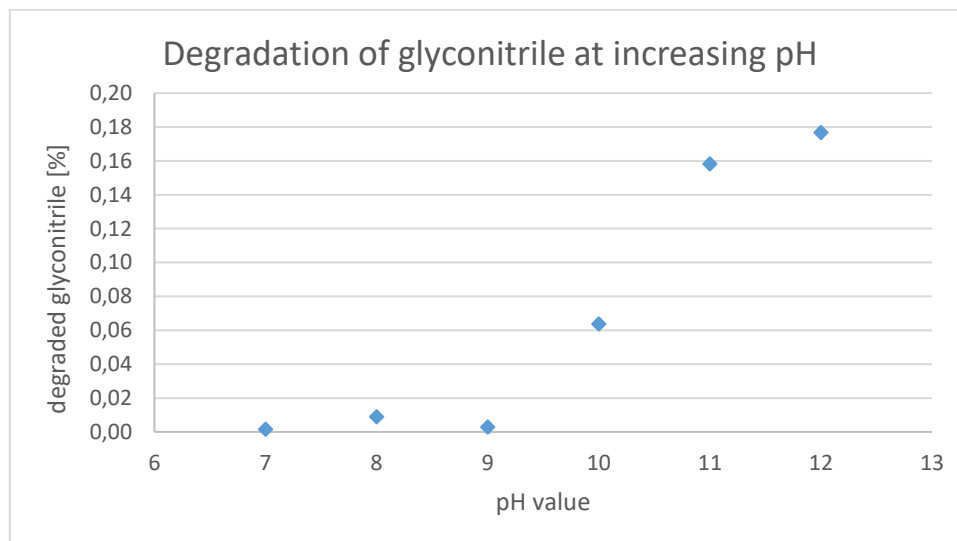


Figure 18: Degradation of glyconitrile at increasing pH; SD smaller than dot

At this point should be mentioned, that it was extensively investigated how the sample pH influences the measurement results of the CNSolution 3000 (see section 7.2). From these experiments it was concluded, that the pH adjustment during the sample preparation itself has no effect on the signal at the detector. That means that samples, which contain the same amount of cyanide, will always produce the same measurement result. This is at least valid in the range of pH 6 to pH 13. Effects that lead to corruption of the measured concentration will also be explained later.

6.2. Reaction time

The reaction time is also a very important factor for the glyconitrile degradation. From the literature is known¹¹, that the equilibrium of the formation reaction of glyconitrile is reached in about 10 minutes. This is also more or less the time between the adjustment of the pH during sample preparation and the actual analysis. However, it is unknown, what happens after this short period of time.

Procedure

Standards with a glyconitrile concentration of ~0.44 mg/L were repeatedly measured at pH 7, 11 and 12 over a duration of at least one hour.

Results and Interpretation

For these three pH values, no change in the amount of released cyanide could be observed (measurement results pH 7: 0.003mg/L CN; pH 11: 0.189 ± 0.005 ; pH 12: 0.183 ± 0.001 ; RSD <1%). As expected, the equilibrium had been reached before the start of the measurement. This is particularly important, because in case of a multi-sample-measurement, the residence time on the autosampler and therefore the time for the formation of cyanide differs significantly.

6.3. Comparison ÖNORM M 6258 – CNSolution 3000

The aim of this experiment was to compare the ÖNORM method to the one used with the CNSolution 3000 (at pH 7 and 11). The standards were produced using waste water, which contained only traces of cyanide by itself. This was done to include matrix effects in the experiment.

To proof the stability of glyconitrile during the ÖNORM M 6258 analysis, the stripped sample was further analyzed with the CNSolution 3000 at pH 11.

Procedure

The complete experiment was done with two identical standards with 0.1 mg/L cyanide and 0.22 mg/L glyconitrile (equals additional 0.1 mg/L cyanide at 100% decomposition; the maximal cyanide concentration assuming 100 % decomposition was consequently 0.2mg/L). As sample matrix, a typical exhaust gas scrubber wastewater was used instead of DI water.

CNSolution 3000: After the production, these standards had a pH close to 7. Hence, no adjusting was necessary for the measurement at pH 7. For pH 11, samples were adjusted with 1M sodium hydroxide solution.

ÖNORM M 6258: Another aliquote of the samples was immediately treated following the standard ÖNORM M 6285⁴ procedure.

Furthermore, the residue of the ÖNORM analysis (content of the flask) was analyzed at pH 11 with the CNSolution 3000. To overcome the buffer capacity of the ÖNORM M 6258 reagents, 5M sodium hydroxide solution was necessary though the fine-adjustment of the pH was done with 1M NaOH.

Results and Interpretation

The following table contains the measurement results, which were corrected by the cyanide concentration of the matrix.

Higher cyanide levels were detected at pH 11 than at pH 7. This result is in accordance with the Cyanohydrin hypothesis. At higher pH, about half of glyconitrile was decomposed and detected as cyanide.

Table 12: Comparison CNS3000 - ÖNROM M 6258 - Experimental results; n=3; SD not shown for clarity;

Method	Sample 1 corr. [mg/L CN]	Sample 2 corr. [mg/L CN]
CNS 3000 pH 7	0.10	0.10
CNS 3000 pH 11	0.15	0.15
ÖNORM pH 3.9	0.07	0.08
ÖNORM residue CNS 3000 pH 11	0.07	0.07

The ÖNORM M 6258 analysis showed values below 0.1 mg/L, which could be explained by the complex procedure. The loss of some of cyanide in the cause of this

procedure can always happen and clearly contributes to the high variation coefficient of this method. For this experiment, it was important to prove, that glyconitrile is not released from the sample. When the residue of the cyanide release was analyzed with the CNSolution 3000, 0.7 mg/L cyanide could be found. This means that nearly all of the bound cyanide remained as such, because it is known from earlier experiments that only 80% of the glyconitrile will degrade at pH 11.

6.4. Conclusion

From the presented experiments using synthetic glyconitrile standards, the previously proposed cyanohydrin-hypothesis could be **confirmed** and additional insight on the properties of this substance was gained.

It was observed, that the pKs of hydrocyanic acid is clearly reflected in the equilibrium between glyconitrile and free cyanide. Because of this, measurements at pH 11 will not only quantify WAD cyanide but also some of the cyanide bound in glyconitrile.

The equilibrium of the decomposition reaction of glyconitrile was reached before the measurement started, and did also not change in a reasonable amount of time. This will be particularly important for the development of an appropriate storage method.

It was also possible to show, that glyconitrile **is not decomposed at a pH of 3.9** (ÖNORM M 6258). It remains in solution during the release of cyanide. Nevertheless, if it is further analyzed using the CNSolution 3000 it could be detected as cyanide.

The following figure shows a graphical representation of the **relationship between released cyanide and the pH value**.

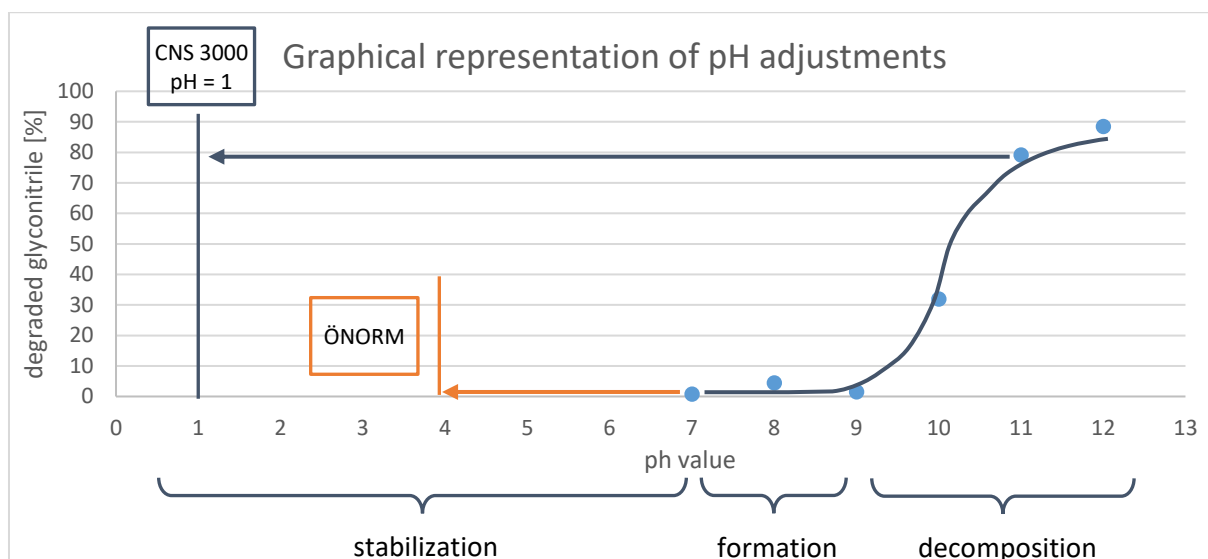


Figure 19: Graphical representation of pH adjustments during analysis with ÖNORM M 6258 and CNS3000; SD smaller than dot

Samples arriving in the laboratory for analysis generally had a pH of approximately 7. The orange arrow indicates the pH adjustment to 3.9 (vertical orange arrow), which is done during the **ÖNORM M 6258** analysis. In this case, no cyanide is released from glyconitrile, because the sample pH never had been in the decomposition region. If the measurement is done according to **OIA-1677**, the pH is raised to 11 during the sample preparation. This is represented by the blue curve, which is defined by the measurement points found by the experiment “pH-dependence of formation and reverse reaction” (section 6.1). In the CNSolution 3000 itself, the pH is lowered rapidly to one (vertical blue arrow). There is not enough time for the reformation of glyconitrile, because at low pH values, no base catalysis is possible. At the detector, the released cyanide is then measured **in addition** to the cyanide originally contained in the sample. The same principle can also be applied when samples are stabilized for storage. If the pH is raised above 9, glyconitrile will become problematic for any kind of cyanide determination.

7. Elimination of the glyconitrile-interference

It had been proven, that glyconitrile leads to **severe interference**, if the pH of the sample exceeds 9 at any time during cyanide analysis. Because this is the case during analysis following OIA-1677³, a new sample pretreatment had to be developed.

7.1. Chemical treatment: ethylene diamine

As already stated (section 5.2), the treatment with ethylene diamine is a suggested method for aldehyde containing samples^{3 5}. However, the underlying mechanism is not explained in literature.

According to the literature, the treatment with ethylene diamine can only prevent further formation of cyanohydrin¹⁷. The regeneration of cyanide is not possible. This information was considered important, as it was thought, that cyanohydrins could somehow be eliminated by this treatment.

However, a closer look on the reaction mechanism revealed that this is not the case. The following figure shows the first steps of the Strecker-Synthesis¹⁹, which works with primary and secondary amines as well.

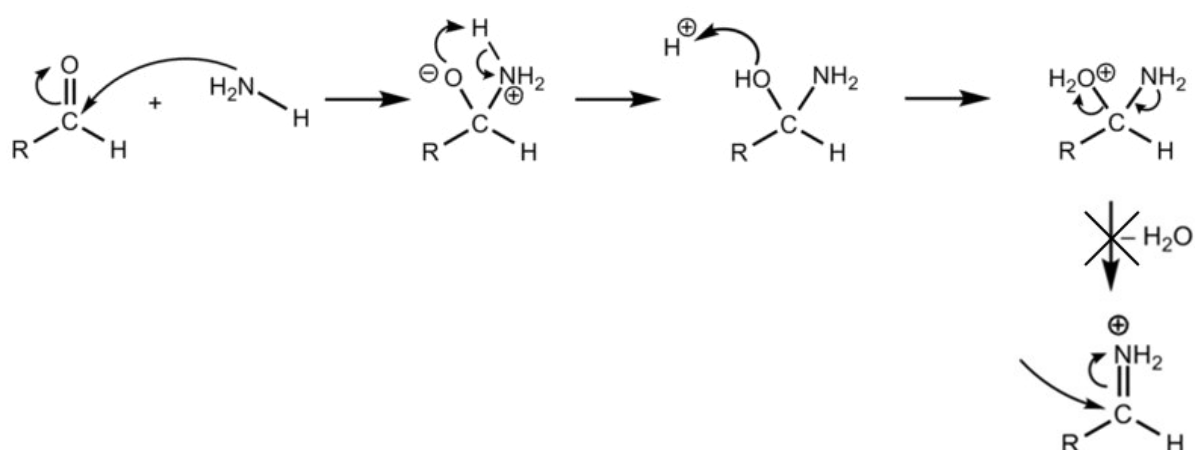


Figure 20: Reaction mechanism ethylenediamine treatment (derived from Strecker synthesis); modified version²⁰

It is evident, that the amine binds to the carbon of the aldehyde group to form a compound that has an amino and a hydroxyl group in geminal position. This is the final stage, if the reaction is performed in an aqueous medium. The elimination of water is not thermodynamically reasonable in this case, though in non-aqueous media it will take place.

Conclusion

The ethylene diamine treatment can only **remove free formaldehyde** from samples. The contained amount of cyanohydrins is not altered and therefore no improvement of the issues investigated in this work were expected.

7.2. Change of measurement pH value

Another promising approach was to avoid raising the pH to 11 before the measurement with the CNSolution 3000. Thereby the decomposition of glyconitrile to cyanide and formaldehyde could be circumvented in a very simple way, though it had to be investigated, if the sample pH itself has any influence on the measurement result.

7.2.1. Description according to the cyanohydrin theory

If the pH of the samples is **not increased to 11** before the measurement, **no decomposition** of cyanohydrin will occur. In this case, a measurement with the CNSolution 3000 should be in better agreement with the ÖNORM M 6258. The green arrow in the following figure represents this new approach.

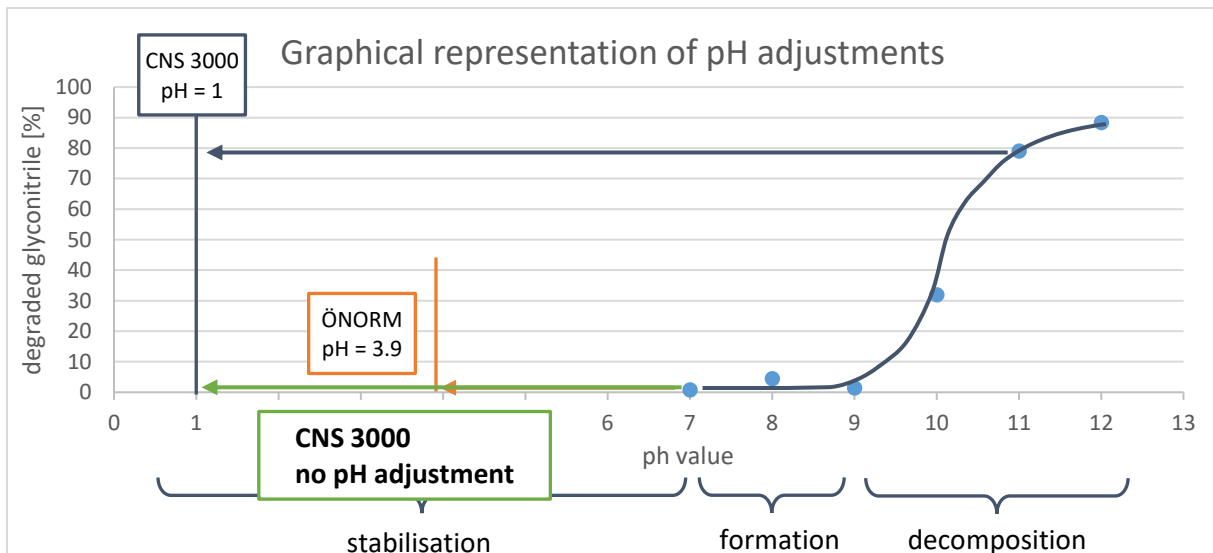


Figure 21: Graphical representation of the direct measurement with the CNS 3000; Updated version of Figure 19; SD smaller than dot

It can be seen, that the pH is **never raised** above the initial sample pH, which represents the same way in which samples are treated during an ÖNROM analysis. Thereby the cyanide level remains constant and in theory, glyconitrile should not cause elevated cyanide data.

7.2.2. Equality of measurements at pH 7 and pH 11

At first, it had to be investigated, if the change of the sample pH has any effect on the measurement with the CNSolution 3000. This is particularly important, because investigations, which were done before this work, indicated some kind of pH dependence of potassium cyanide standards if they are adjusted to a lower pH than 12.

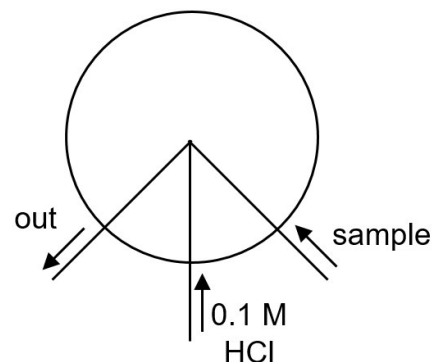


Figure 22: Mixing chamber CNS 3000

Theoretical investigation: pH value in the diffusion chamber

Because the setup was explained in detail in section 2.2, only the mixing chamber, which is positioned between the sample loop and the diffusion chamber, will be depicted schematically here (Figure 22).

To determine the pH of the outgoing stream, the flow rates and the pH of the entering streams had to be considered. The sample still had a pH of 11 at this position and the pH of a solution containing 0.1M hydrochloric acid was one. The flow rates were unknown, and therefore had to be measured experimentally.

Flow rate measurement

The tubes for HCl and sample were disconnected from the mixing cell and the amount of liquid emitted in 60 seconds was collected in a tared beaker. In the case of the out-stream, the collection is done at the entrance to the diffusion cell. The following table contains the results for all three streams.

Table 13: Flow rates of the mixing chamber streams

stream	Flow rate [g/min]
HCl	0.92±0.04
sample	0.86±0.03

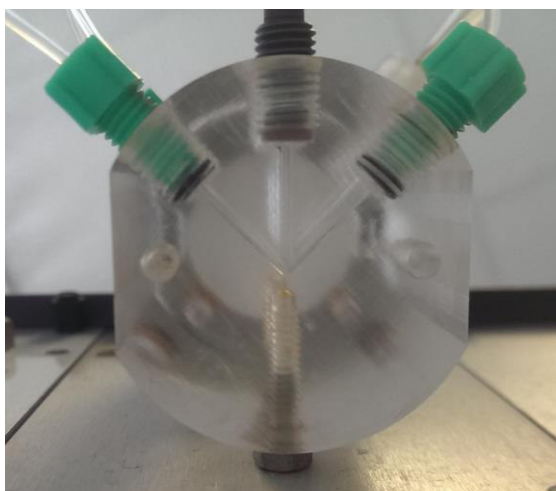


Figure 23: Picture of the mixing chamber (bottom up view)

The flow rate of the inlets were nearly the same. The HCl stream might have a little bit lower throughput because it has to pass the sampling unit on its way between the pump and the mixing cell. Due to backpressure in the mixing cell the outgoing stream had a lower flow rate than the incoming ones, if they were not connected to the mixing cell. Due to this tailback, it was concluded that the data on the flowrates only had **qualitative** character. Because it was known that, the

channels of the mixing chamber (Figure 23) have the same diameter, the mixing ratio of the incoming stream should rather be **1+1**.

Calculation of the pH in the gas diffusion chamber

The samples are acidified, because only hydrocyanic acid, the protonated form of cyanide, can pass the hydrophobic membrane in the gas diffusion chamber. The following table contains the calculated pH values present in this component for samples with different pH. The results are based on a 1+1 mixing of 0.1 M HCl solution and the different samples.

Table 14: pH in the diffusion chamber for different samples

Sample type (pH)	Calculated pH in the diffusion chamber
Calibration standard (pH 12)	1.04
Regular sample (pH 11)	1.0043
Stabilized sample (pH 8)	1.0000043
Untreated sample (pH 7)	1

Conclusion

Because this is a **theoretical comparison**, the significance of the results was not considered. From a theoretical point of view, **no pH dependence** should be expected pH range from 7 to 12. The pH at the gas diffusion membrane is in any case **below the pKs of hydrocyanic acid** (9.4). The origin of this issue is the focus of subsequent investigations.

7.2.3. Measurement of standards with pH values lower than 11

To prove equality of the measurement at pH 7 and pH 11 it was necessary to produce standards from potassium cyanide and potassium tetracyanozincate. It was observed that synthetic standards show some kind of pH dependency, if their pH is adjusted to lower values. A series of experiments was done to demonstrate, that the instrument did not cause this issues. Contrary to first assumptions, it was suspected that the so-called **CN-loss** (short for cyanide-loss) was caused by the pH adjustment with very small amounts of acid itself.

1. CN-loss pH 7 – pH 9.4 – pH 11

To show that low spike recoveries were not caused by the instrument or the method used, a standard adjusted to pH 7 was measured at pH 7 (sample 1) and at pH 11 (sample 2). Further, another part of the same standard was measured at its production pH of 9.3 (sample 3) and afterwards also at pH 11 (sample 4).

This will allow a clear distinction between the two possible scenarios potentially responsible for cyanide loss:

1. A pH dependence of processes within the **instrument** could cause the reduced measurement signal. In this case, there would be different measurement results for sample 1 and 2 and also for 3 and 4. However, sample 2 and 4 should show the same result.
2. The reduced measurement signal might also be induced by the **pH adjustment** itself. This process involves the actual loss of cyanide during the preparation of the sample. This can be expected to lead to the same result for the measurements of the samples 1 and 2 and for the samples 3 and 4. But more important, the cyanide concentration of the samples 3 and 4 will be close to the ideal value, whereas the first two samples will have a lower cyanide content.

Procedure:

An unstabilized 0.2mg/L Cyanide standard (sample 3) was produced from solid potassium cyanide. As a reference, one part was immediately adjusted to pH 11 (sample 4). Another 200 mL of the standard were transferred to a beaker and treated with a few drops of very diluted HCl to lower its pH to 7. Afterwards it is divided into two beakers, which were covered and left at the lab bench until just before the beginning of the measurement (sample 1 and 2). At the same time as the autosampler takes up sample 1, the content of the one beaker was brought to pH11 (sample 2). To make this procedure more convenient, the samples were positioned on the autosampler in the following order: 4 -> 1 -> 2 -> 3. The complete experiment was repeated twice.

Results and Interpretation

Table 15: Results CN-loss pH 7 – pH 9.4 – pH 11; max. standard deviations = 0.002mg/L; n=3; SD not shown for clarity;

Sample/experiment	1	2
	[mg/L] (recovery)	[mg/L] (recovery)
Sample 1 (pH 7)	0.180 (89.9%)	0.175 (87.5%)
Sample 2 (pH 11)	0.186 (93.1%)	0.169 (84.5%)
Sample 3 (pH 9.3)	0.204 (102%)	0.208 (104%)
Sample 4 (pH 11)	0.200 (100%)	0.208 (104%)

Sample 3 and 4 always showed the desired value of 0.20mg/L. For sample 1 and 2 the cyanide recovery were lower (80-90%), but no significant difference could be seen between the measurement at pH 7 and 11. This leads to the assumption, that the previously described **case two** could be applied here.

This implied that the measurement procedure itself is **not pH dependent, but the sample preparation causes the observed losses**. The amount of cyanide, which is contained in the sample, will always be determined correctly in the range between pH 7 and 12. But if the pH is adjusted by addition of acid, the cyanide recovery is lowered by approximately 10 to 15%.

2. CN-loss

As already demonstrated, the pH adjustment had significant influences on the cyanide recovery in synthetic solutions. A series of experiments was conducted to further investigate the extent of the CN-loss. In this section, the most significant results are presented.

pH dependence

Procedure: A series of cyanide standards containing 0.2 mg/L CN (pH 9.2) was set to **pH 7, 8 and 9** using strongly diluted hydrochloric acid. Immediately ($t < 5\text{sec}$) after this initial pH adjustment, the pH of these standards was readjusted to 11.

Because of the low buffer capacity of the unstabilized standards only one drop of 1M NaOH had to be added to obtain a **pH of 10**. To ensure consistent test conditions, this standard was given the same residence time ($t < 5\text{sec}$) as the first three samples. After this, the pH was adjusted to 11 as well.

The pH of the last standard was only raised to **pH 11**, without any other treatment. This procedure was selected to simulate the loss of cyanide when unstabilized standards are adjusted to different pH values. It was also attempted, that all standards had the same residence time ($t < 5\text{sec}$) at a certain pH value.

Results and Interpretation: Lowered cyanide recovery was found for samples, which were adjusted to a lower pH. In contrast to this, the sample, which was only treated with NaOH, showed a recovery rate of nearly 100%. The following figure shows all results.

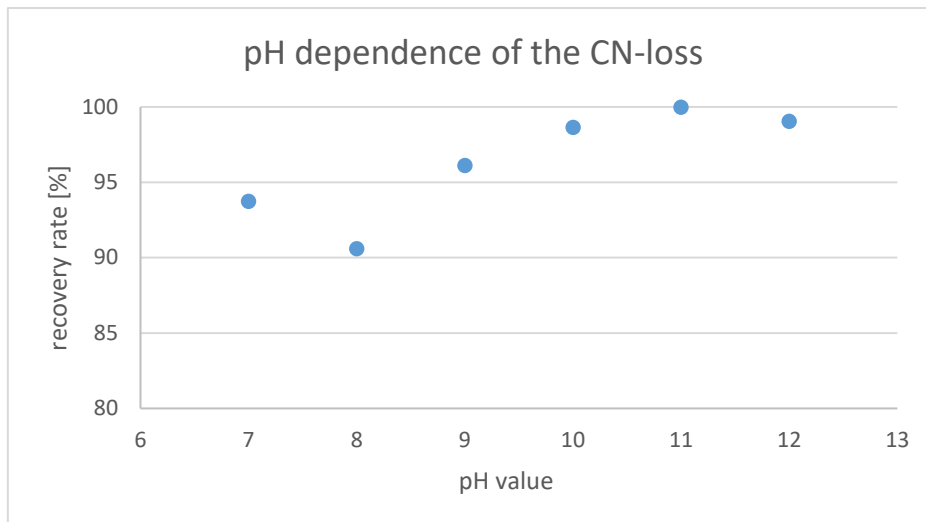


Figure 24: CN-loss during the adjustment to different pH values;; n=3; SD smaller than dot

Time dependence

Procedure: 500mL unstabilized 0.2mg/L CN standard (pH 9.3) were produced (see above). Each of seven 100 mL beakers was filled with approximately 70mL of this standard and the pH was lowered to 7 with strongly diluted HCl. The individual samples were left uncovered at the lab bench for 0, 1, 2, 5, 7, 10 and 15 minutes. Afterwards the pH of all samples was raised to 11, which prevented further release of cyanide from the solution. A reference sample was produced by setting the pH to 11 immediately after production.

Results and Interpretation: The reference sample showed exactly the expected value of 0.2mg/L cyanide. Systematic errors from the production of the standard (e.g. minute dilution of the sample by pH adjustment) could therefore be excluded.

It is apparent, that the pH adjustment itself was responsible for a CN-loss of approximately 8%. More cyanide was not lost until 10 minutes passed by. The data are shown in the following figure.

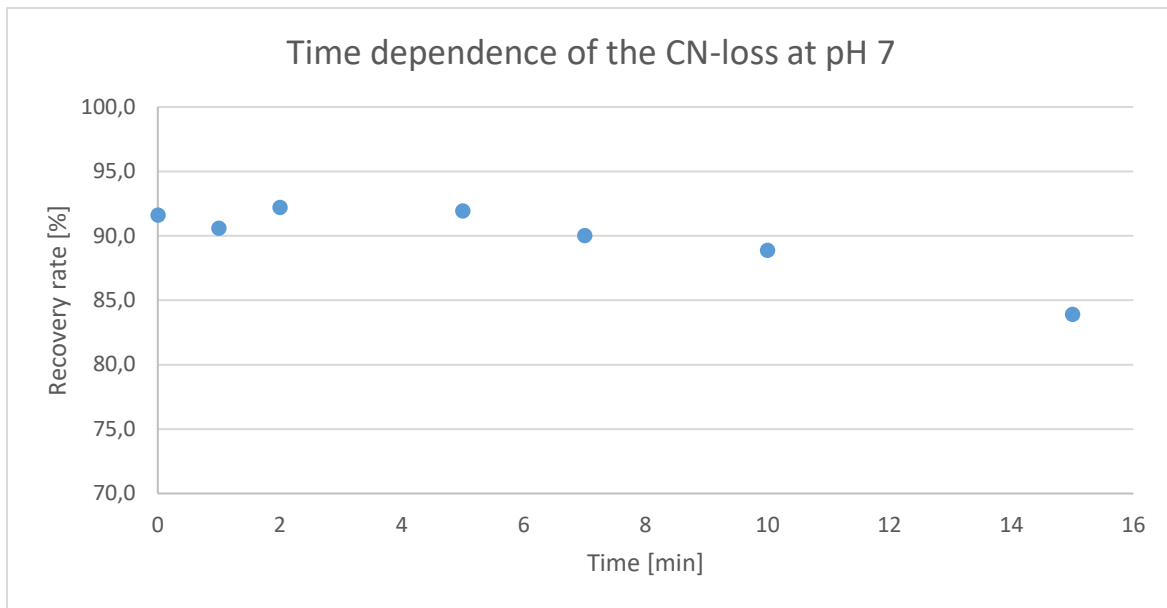


Figure 25: Time dependence of the CN-loss at pH 7; n=3; SD smaller than dot

3. Consequences for ÖNORM M 6258⁴

After evaluation of these experiments, it became clear, that the CN-loss may also be an issue during the analysis following ÖNORM M 6258. During the sample preparation, the pH value is lowered to 3.9 in uncovered beakers. The transfer to the release apparatus does not happen until all samples are prepared, as discussed in the introduction.

This sample preparation step deviates from the original procedure described by ÖNORM M 6258, that recommends to perform the whole sample preparation in the three-necked flask of the release apparatus while the gas flow through the absorption solution is already established.

Although the modified version of ÖNORM M 6258 had been validated in the past, additional investigations are needed.

7.2.4. Validation experiments

In this section, it will be shown that it is legitimate to change the measurement pH of OIA-1677 from **11 to 7**. This will be done by a series of comparative experiments that are evaluated statistically.

1. Comparison of calibration curves at pH 7, pH 11 and pH 12

To show, that the CNSolution 3000 is usable at pH 7 as well as at pH 11, calibration curves were recorded at both of these pH values. The investigated concentration range was between 0.02 and 0.2 mg/L cyanide.

Procedure:

The standards were diluted from an unstabilized KCN stock solution and acidified to pH 7 with 3 to 4 drops of highly diluted hydrochloric acid. Until the measurement starts, the samples were stored in the same closed graduated flasks, which were used for their production. The transfer to the autosampler happened not until moments before the measurement starts. At the same time, the unused residual of the sample was stabilized with 2 drops of 1M NaOH solution. After all pH 7 standards had been analyzed, the residuals were adjusted to pH 11 and afterwards measured themselves (CN-loss compensation).

Evaluation:

Before each set of measurements a calibration is necessary, because the electrical signal (pA) can drift within several hours. Because of this raw signals cannot be compared directly. At first, they had to be converted to concentrations.

The goal of this experiment was to show that the deviation between the measurements at pH 7 and pH 11 was lower than the measurement uncertainty (twice of the standard deviation of the calibration). Therefore, the highest value of this sample series was used. The software-based automatic elimination of outliers was not taken into account. Additionally, the calibrations were evaluated using the software "Validata".

Results:

Measurement uncertainty: The calibration of the measurement “0922_03” (see appendix) had a standard deviation of 2%. Therefore, the two-sigma boundary for this method was **4%**, which was seen as a very reasonable value. As it was mentioned before, this was far better than the ÖNORM M 6258⁴ method. The relative standard deviation of all results for this method in a ring trail was 28%. Although ring trails have a rather high RSD compared to single measurements (different operators, instruments, days, etc.), it is clear that by using the CNS3000 a higher method precision can be attained.

The following table contains the results for the measurements at pH 7 and pH 11. In both cases. The calibration was done at pH 12.

Table 16: Results of the comparison of calibration curves at pH 7 and pH 11; ; n=3; SD not shown for clarity;

Standard	pH 7 [mg/L CN]	pH 11 [mg/L CN]	pH 11 – pH 7 [mg/L CN]	Delta [%]
1	0.017	0.017	0.000	-0.22
2	0.043	0.044	0.001	2.96
3	0.092	0.089	-0.003	-3.37
4	0.138	0.135	-0.003	-2.11
5	0.184	0.183	-0.001	-0.78

Validata (Excel macro for method validation): It is known that the measurement can be done correctly at pH 11. For this reason, pH 7 standards were also measured at pH 11. By doing so, the “true” concentration of the standards could be determined. It was lower than the concentration, which would have been achieved only by dilution of the stock solution, because of the pH adjustment. During the “Validata” analysis, the concentration values determined at pH 11 were combined with the peak heights (in pA) from the measurement at pH 7.

To allow a comparison, a calibration curve at pH 12 was also recorded.

pH 7: Between 0.02 mg/L and 0.2 mg/L, **variance inhomogeneity** could be observed. Consequently, the operating range was reduced to 0.05 to 0.2 mg/L.

pH 12: Variance inhomogeneity could also be observed when standards with pH 12 were measured. Again, the elimination of the calibration point with the lowest concentration could solve this problem.

Linearity and the operating range **are secured** for pH 7 and 11.

The complete validate reports can be seen in the appendix section 11.3.

A **F-test** was performed to compare the method standard deviation. Therefore the concentration determined at pH 11 was used for both series (pH 7 and pH 11). The calculated test value (PG=1,52) was smaller than the tabulated one (PG=4,11; confidence 90%). The recovery function further showed that **no proportional or constant systematic deviation**. This indicated that both measurements are equal.

2. PbCO₃ treatment

In order to remove sulfide (severe positive interference) from samples, a treatment with PbCO₃ is recommended by OIA-1677³. It had to be investigated if this method works at a pH of **7** in a concentration range between 0.02 and 0.2 mg/L. Further, it was investigated how a **variation of the amount** (0.1g-2g) of PbCO₃ influences the measurement.

2.1. 0.02 – 0.2 mg/L Cyanide + 0.2 g PbCO₃/100mL

100 mL Cyanide standards with a concentration of 0.02 – 0.2 mg/L were treated with 0.2 g PbCO₃. The method OIA-1677 recommended to add PbCO₃ **immediately after the sample collection**. A pH value of 7 should therefore be no problem. It had to be proven that no problems occur during the measurement of PbCO₃ containing samples.

Procedure:

The standards were produced from unstabilized potassium cyanide stock solutions. For the acidification to pH 7, highly diluted hydrochloric acid was used. The standards were left covered on the lab bench until one minute before the analysis. Then they were transferred to PbCO₃ containing beakers. Turbidity could be observed. For the transfer to the auto sampler, syringe filters (Rotilabo 25mm, 0.45µm pore size, CME membrane, PVC housing) were necessary.

The residuals of the samples were stabilized with two drops of a 1M sodium hydroxide solution and covered with “Parafilm”. The exact adjustment to and the measurement at pH 11 was done after the measurement at pH 7.

Results:

The following table contains the results of the measurements at pH 7 and pH 11.

Table 17: Results 0.02 – 0.2 mg/L Cyanide + 0.2 g PbCO₃/100mL; mean values, rounded; n=3; SD not shown for clarity;

Standard	pH 7 [mg/L CN]	pH 11 [mg/L CN]	Delta [%]
1	0.018	0.018	-2.0
2	0.045	0.045	-0.8
3	0.093	0.093	-0.4
4	0.139	0.137	-1.6
5	0.184	0.188	1.8

The deviation between pH 7 and pH 11 was in all cases small than the precision of the method of 4%. This indicated that regardless of the pH value is 7 or 11 the method works **equally well** in the presence of PbCO₃.

The evaluation of calibration data using “Validata” showed **variance inhomogeneity** again. The highest concentration level was eliminated. Additionally, the linearity test was not passed. This can be explained by a slight deviation of the calibration points 0.15mg/L and 0.20 mg/L from the calibration curve. The analysis of residuals, which was plotted by validata can be seen in Figure 26. For the complete validata report, see appendix section 0.

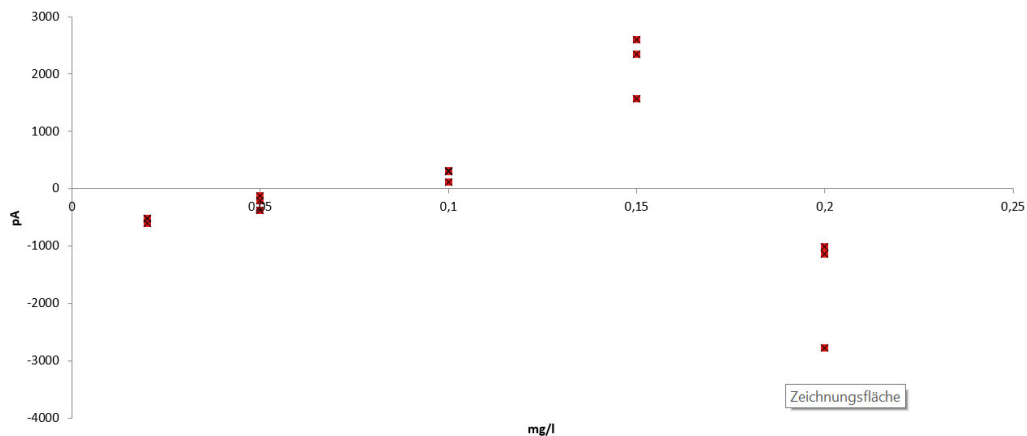


Figure 26: Analysis of residuals of 0.02 – 0.2 mg/L Cyanide + 0.2 g PbCO₃/100mL from validata

The pH 7 and the pH 11 methods were also compared using the F-test, which showed that these two techniques are **equal** for a confidence of 90%. The analysis of the recovery function showed, that **no constant or proportional systematical deviation** is expected (confidence 95%)

2.2. PbCO₃ – excess and shortage

In routine analysis, the amount of PbCO₃ was not weight exactly before the application. It was recommended to use one “**spatula tip**”. To examine the effect of variations in PbCO₃ addition, different amounts of PbCO₃ were tested.

Procedure:

The standards were produced from an unstabilized KCN stock solution, which was adjusted to pH 7 with highly diluted hydrochloric acid. The samples were stored in volumetric flasks until one minute before the analysis. Then the standards were transferred to beakers, which contained different amounts of PbCO₃. To avoid contaminations of the instrument, syringe filters (Rotilabo 25mm, 0.45µm pore size, CME membrane, PVC housing) were used during autosampler loading. Again, the remains of the samples were stabilized with two drops of sodium hydroxide solution. The exact adjustment was done after the first determination was finished.

Results:

The following table contains the results of the samples, which were treated with different amounts of PbCO_3 and determined at pH 7 and pH 11.

Table 18: Results of the PbCO_3 variation; data for 0.2g PbCO_3 from previous experiment 2.1.; n=3; SD not shown for clarity;

PbCO₃ [g]	pH 7 [mg/L CN]	pH 11 [mg/L CN]	delta [%]
0.1	0.151	0.149	-0.9
0.2	0.184	0.188	1.8
0.5	0.165	0.166	0.8
1	0.185	0.183	-0.9
2	0.188	0.186	-1.2

No **significant difference** could be observed between the measurements at pH 11 and pH 12. The maximum deviation is **1.8%**.

A shortage or excess of PbCO_3 should therefore **not cause any problems** if no sulfide is present. Nevertheless, the reasonable amount of **0.2g**, which is approximately one spatula tip, is recommended.



Figure 27: 0.2g of PbCO_3

3. Cyanohydrin – Dilution into the operation range

During routine operation, situations of **high cyanide concentrations** can occur. Usually the deviation between ÖNORM M 6258 and OIA-1677 is particularly high in this case. To make samples analyzable, they have to be **diluted** into the operation range (0.02-0.2 mg/L). This experiment should show that this is applicable for samples, which contain cyanide as well as glyconitrile.

Procedure:

The standards were produced from unstabilized cyanide stock solutions. At first, the volumetric flasks were filled halfway with DI water. Then the necessary amount of cyanide and glyconitrile stock solution (~577g/L; Sigma Aldrich 374768²¹; lot#: BCBF4937V; CAS-number 107-16-4) was added. After that, the flasks were filled up to the mark. The molar ratio between these two compounds is 1:1 at all concentrations. Table 19 shows the concentrations of glyconitrile and cyanide present in the standards.

Table 19: Cyanide and glyconitrile concentration of standard 1-5

# Standard	c(CN) [mg/L]	c(Cyh.) [mg/L]	Dilution
1	0.10	0.21925	-
2	1	2.1925	1+9
3	10	21.925	1+99
4	20	43.85	1+199

The dilution into the operation range was done according to the factors which were given in the previous table. The pH of the samples 1 and 4 were 8.24 and 9.84, respectively. The adjustment to pH 7 was done with highly diluted hydrochloric acid. The samples were stored in volumetric flasks until the analysis started. Again, the samples were stabilized with two drops of 1M NaOH solution and adjusted to pH 11.2 after the first measurement was finished.

NOTE: It has to be considered, that the CN-loss **cannot be determined** during this experiment. For any other measurement in this section, at first a measurement was done at pH 7. Then the same samples were also determined at pH 11 to obtain comparable results. However, if glyconitrile is also present in the system, the cyanide concentration will rise with increasing pH value.

From previously experiments (section 7.2.3), it was assumed that the CN-loss was also in the range of 10-15%. In any case, the cyanide concentrations **should never rise** above the initial level.

Results:

Table 20: Results Validation experiment 3 - Dilution into the operation range; n=3; SD not shown for clarity;

Standard [mg/L CN]	Measured conc. pH 7 [mg/L]	Deviation from ideal value [mg/L]	Measured conc. pH 11 [mg/L]	Deviation ph 11- pH 7 [mg/l]
0,1	0,093	0,007 (7%)	0,174	0,081
1	0,100	0,000 (0%)	0,179	0,079
10	0,097	0,003 (3%)	0,178	0,081
20	0,096	0,004 (4%)	0,176	0,081

All of the measured standards showed results **close to the ideal value**. It seems that glyconitrile-cyanide-systems have a lower CN-loss, because at pH 7 nearly all of the cyanide can be recovered.

It can also be seen that at pH 11 the same amount of glyconitrile is degraded in all samples. The amount is also in the range of the previous experiments.

4. Conclusion of the Validation experiments

In experiment one, it was shown that the measurements of the same samples at pH 7 and pH 11 lead to **similar results**. Deviations between them are within the uncertainty

of the method and are nearly an order of magnitude **lower** than at the ÖNORM M 6258⁴.

It was also shown that samples treated with PbCO_3 could be measured at pH 7. The amount of PbCO_3 should be approximately 0.2g, but there are no negative effects encountered up to 2g.

No problems will also arise from samples with a cyanide content up to **20mg/L** as dilution works **perfectly fine** to obtain samples within the calibration range of the method.

7.2.5.Storage of samples before and during the measurement

Below its pKs value (9.4), the majority of cyanide is obviously present as **hydrogen cyanide**. This **volatile** compound is able to evaporate from the liquid phase and is consequently lost. Because the newly developed method involves the measurement at pH 7, it is considerably likely that lower results will be found. The following experiments were done to estimate the magnitude of this effect.

1. CN-loss during the measurement

As described before, cyanide is lost to the gas phase from uncovered samples at pH 7. In this experiment, it was examined how the cyanide concentration of a real sample changes, if it was left uncovered for 1.5 hours on the autosampler.

The only sample treatment was filtering. Three test tubes were filled and measured five to six times each. The time dependent CN-loss of a real sample is described by 16 mean values plotted in Figure 28.

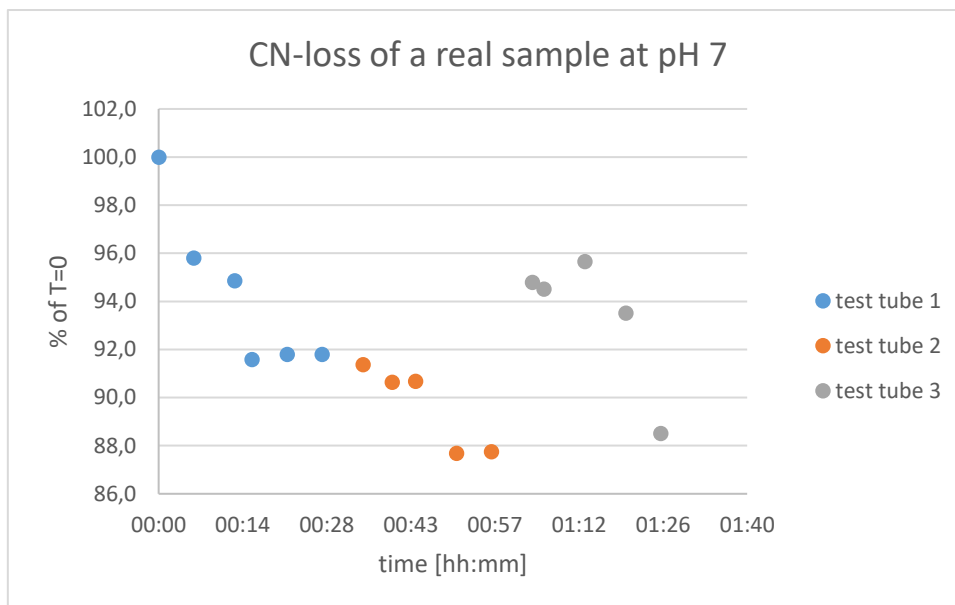


Figure 28: CN-loss of a real sample on the autosampler; three test tubes containing with the same standard

From Figure 28 it can be deduced that in general the cyanide concentration continuously decreased during the first hour. With the cyanide loss in test tube 3 however doesn't follow this general trend, though the reason for this remains unclear. However, the generalized statement appears valid, that the cyanide loss in samples of pH 7 involves several factors besides time. The nature of these additional factors remains unclear and further investigations will be needed to fully understand the cyanide loss at pH 7.

2. Sample storage at pH 6.5

As already discussed sample storage at high pH values should be avoided, because of the decomposition of glyconitrile. However, it is known that cyanide is lost to the gas phase at medium to low pH values. This experiment should clarify if storage in **bubble free sealed flasks** in a **refrigerator** is a viable solution to this problem.



Figure 29: Bubblefree cooled storage using Erlenmeyer flasks and polymere plugs

In routine analysis, samples with a pH value of approximately 6.5 had been observed in some cases. Therefore, samples with this pH were plausible to be investigated during this work.

Procedure: Four samples were produced from an unstabilized potassium cyanide stock solution. The pH adjustment was done with highly diluted hydrochloric acid. Each sample was transferred to a 300ml ground joint Erlenmeyer flask, which was filled until it overflows. Bubble free sealing was achieved by application of a polymer plug, while the flasks were slightly tilted. At the same time, a reference sample was taken from each standard to determine the cyanide concentration before storage.

Sample 1 and 2 remained bubble free sealed in the refrigerator for 24 hours. After the first determination, they remained another day sealed in the fridge but this time with a little air bubble on top. Sample 3 and 4 were analyzed after a storage time of 48 hours. All measurements were done after adjusting the pH to 11 prior analysis.

Results:

Table 21: Sample storage pH 6.5; measured concentrations; n=3; SD not shown for clarity;

sample	c day 0 [mg/L]	c day 1 [mg/L]	c day 2 [mg/L]
1	0.180	0.184	0.188
2	0.188	0.191	0.192
3	0.181	-	0.192
4	0.187	-	0.194

Table 22: Sample storage pH 6.5; concentration drop in mg/L and %

sample	Drop after 24h [mg/L]	Drop after 48h [mg/L]	Drop after 24h [%]	Drop after 48h [%]
1	-0.005	-0.009	-2.6	-4.8
2	-0.003	-0.004	-1.5	-2.0
3	-	-0.011	-	-6.2
4	-	-0.007	-	-3.8

All samples showed a slightly increased cyanide concentration though this can be explained by the **uncertainty** of the measurement. In any case, **no cyanide loss** was

observed. Therefore, this method seems to be a **reasonable alternative** to sample storage at high pH values.

Furthermore, a stabilization with zinc ions should be investigated in the future. The formation of a weak cyano-complex could improve the stability of free cyanide. This technique is already used by Merck²² at their cyanide stock solutions.

8. Formaldehyde analytics

Unfortunately, it was not possible to examine real wastewater samples during this work, because of the **insufficient content** of glyconitrile. The reason was the unavailability of samples with elevated formaldehyde concentrations. Formaldehyde is present in the investigated wastewater samples only during atypical process conditions in the gas scrubber. However, a method to determine the **formaldehyde** concentration is presented. As discussed before (see section 5.3), this aldehyde can be released from samples containing glyconitrile and is therefore seen as an indicator for this interference.

Method description: Formaldehyde-2,4-DNPH HPLC-DAD

A convenient way to determine formaldehyde is **derivatization** and quantification using **HPLC-DAD**. In the first step 2,4-dinitrophenylhydrazine is added to the sample. After the derivative has formed, it is extracted using SPE cartridges.

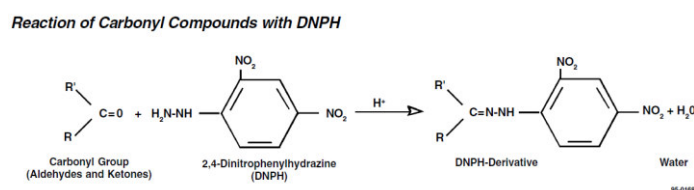


Figure 30: Reaction Equation Aldehyde Derivatization²³

The elution is done with acetonitrile. The high performance liquid chromatography with diode array detection uses a RP 18 column (5 μ m, 4.6 x 150mm) and is operated isocratically (solvent MeOH:H₂O 70:30; flow rate 1mL/min). 360nm is the recommended detection wavelength. The measurable concentration range starts at **1mg/L** is linear at least up to 25mg/L.

This analysis should be performable relatively easy, if suitable instrumentation is available. Otherwise, contract laboratories, like the **Institute of Analytical Chemistry and Food Chemistry** at the **TU Graz**, are able to do this determination.

9. Conclusion

The aim of this work was to identify a **positive interference**, which caused a deviation between the cyanide measurement results of OIA-1677³ and ÖNORM M 6285⁴. Further, an appropriate technique to avoid this should be developed.

In the course of this work, both methods were extensively investigated and characterized (see section 2). The most common interferences were investigated, but none of them could be identified as the source of the observed issues (see section 4).

Glyconitrile was suspected to release cyanide at high pH values. It was observed that the **pH** has a strong impact on the glyconitrile-cyanide-equilibrium. Three regions could be identified. The formation happens at pH 7 to 8, the decomposition above pH 9 and below pH 7 glyconitrile is stable. This is the reason why the cyanide determination following the ÖNROM is not affected by this interference. During its procedure, the pH is never raised above 8 and therefore no glyconitrile decomposes resulting in not additional release of cyanide. (see section 6) Much to the contrary, following OIA-1677 a pH change of the sample to pH 11 is requested by this norm.

However, the **pH adjustment** to higher values during the sample preparation should be avoided as otherwise the glyconitrile-decomposition takes place.

A new method based on OIA-1677 was developed wherein the pH remains at 7 and validation experiments showed that this new method is **statistically coherent** with the old one and can be **used as a replacement** (see section 7).

It was also investigated how the **sample storage** should be done in the future. A new bubble-free and cooled storage method was characterized and showed promising results (see section 7.2.5). Further investigations and longtime studies on this topic and especially on the stabilization using ionic zinc are highly recommended.

Another issue, which came to mind during the investigations, is the potential loss of cyanide to the gas phase during the adjustment of the sample pH value. If the pH is lowered to 7, the cyanide concentration decreases in the range of 10%. Although a

significant impact is not suspected, additional research is recommended, to examine the influences of this effect on the procedure of ÖNORM M 5862.

In section 10 the new **Recommended Procedure** is presented. If glyconitrile is present in samples, its application is highly recommended. It is based on the existing method (OIA-1677) and harmonizes it to the ÖNROM M 6285.

10. Measurement instructions for the pH 7 method

As the last step of this work, a new **Recommended Procedure** for the sample preparation and conservation was developed. It is based on the existing methods OIA-1677³ and ISO-14403⁶ and takes into account the newly gained knowledge about cyanohydrin interference.

Sample Preparation and Conservation

300mL of samples, which are not analyzed immediately, have to be transferred into a 300mL Erlenmeyer flask (with ground joint) and sealed bubble free with a polymer plug. The flask has to be kept in a refrigerator.

50mL of samples, which are analyzed immediately, should be transferred to a 100mL beaker. It should be covered air tight with Parafilm. If the presence of sulfides is suspected, a spatula tip (~0.2g) of PbCO_3 should be added to the samples shortly before the measurement starts. After short stirring the sample is drawn up into a syringe and transferred to an autosampler test tube through a syringe filter (Rotilabo 25mm, 0.45 μm pore size, CME membrane, PVC housing). The first 3mL have to be discarded.

The handling of the instrument itself is the same as described in the CNSolution 3000 operators manual.

11. Appendix

11.1. Additional pictures of CNSolution 3000



Figure 31: CNSolution 3000 diffusion chamber side view

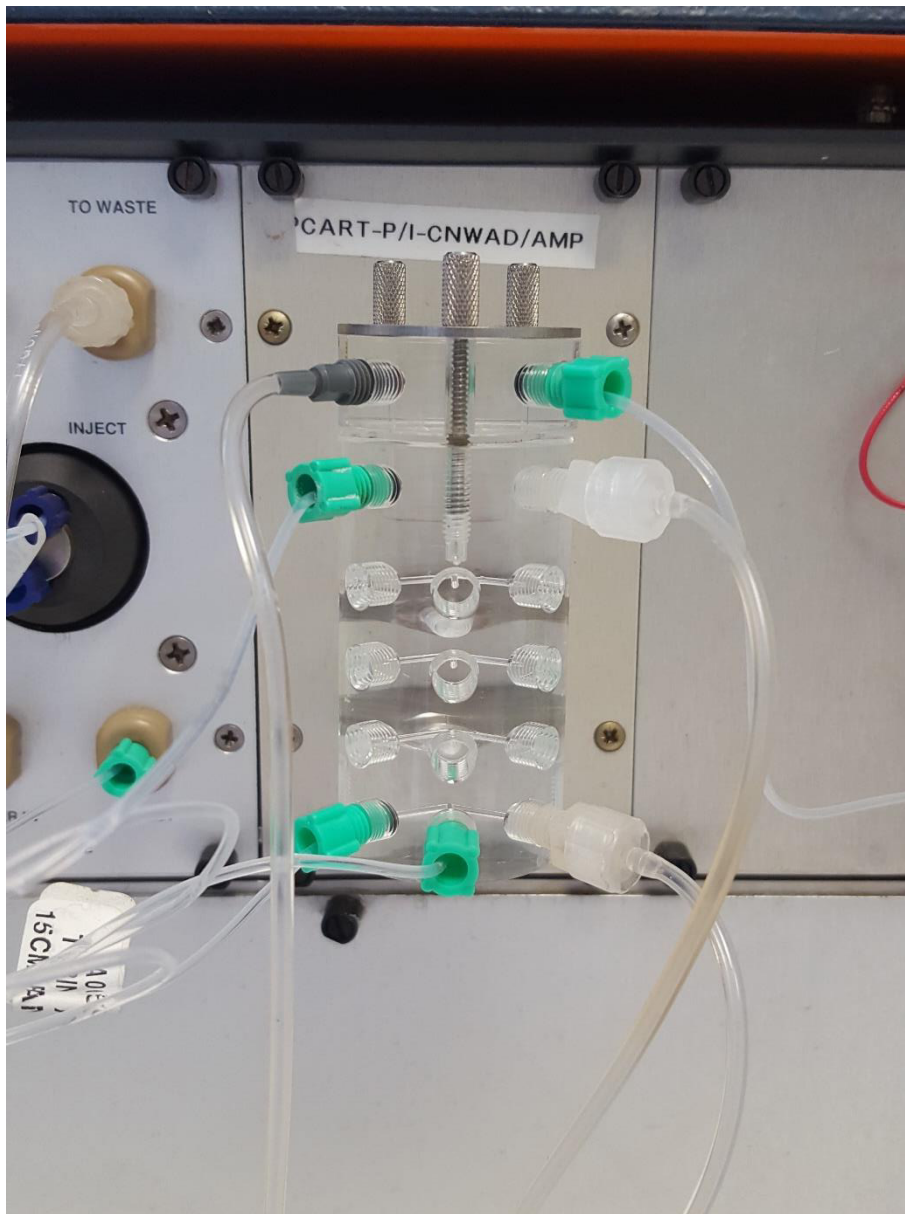


Figure 32: Mixing cell and diffusion chamber module



Figure 33: 120-position autosampler



Figure 34: CNSolution

11.2. Ad CNSolution; examples for relative standard deviation of the calibration

Ergebnisreport des Laufs
 Ergebnisse: L:\LABOR\CN_ANA-1\MOSE-1\REPORTS\0928_04.RST
 Ergebnisse fertiggestellt: 11:32 September 28, 2015.
 Benutzer: moser

V1 *Halogenoide* *pH12*

Zeit	Prob	Name	Höhe	Kalk.	Flagg
----- Benutzer Nachfrage: Datensammlung beginnen -----					
10:47	0	Carryover	628	0.000842	LO
10:47	0	Baseline	0	0.000000	BL
10:48	1	Cal 0.200 mg	144184	0.193293	LO
10:48	1	Cal 0.200 mg	156045	0.209192	LOOL
10:53	1	Cal 0.200 mg	149372	0.200247	LO
10:54	1	Cal 0.200 mg	147150	0.197268	LO
		!Zweck & RSD	146902	0.196936	3.43%
10:54	0	Blank	128	0.000171	LO
10:55	0	Read Baselin	0	0.000000	BL
10:56	2	0,02 pH 12	15109	0.020255	LO
11:00	2	0,02 pH 12	15211	0.020391	LO
11:01	2	0,02 pH 12	15118	0.020267	LO
		!Zweck & RSD	15146	0.020304	.373%
11:03	0	Read Baselin	0	0.000000	BL
11:04	3	0,05 pH 12	37760	0.050620	LO
11:04	3	0,05 pH 12	37973	0.050906	LO
11:05	3	0,05 pH 12	38159	0.051155	LO
		!Zweck & RSD	37964	0.050894	.526%
11:06	0	Read Baselin	0	0.000000	BL
11:09	4	0,10 pH 12	75213	0.100830	LO
11:10	4	0,10 pH 12	75733	0.101527	LO
11:11	4	0,10 pH 12	76271	0.102248	LO
		!Zweck & RSD	75739	0.101535	.699%
11:12	0	Read Baselin	0	0.000000	BL
11:13	5	0,15 pH 12	114971	0.154129	LO
11:16	5	0,15 pH 12	119132	0.159708	LOOL
11:18	5	0,15 pH 12	116177	0.155746	LO
		!Zweck & RSD	115574	0.154937	1.85%
11:18	0	Read Baselin	0	0.000000	BL
11:19	6	0,20 pH 12	155024	0.207824	LO
11:20	6	0,20 pH 12	156737	0.210120	LO
11:22	6	0,20 pH 12	156235	0.209447	LO
		!Zweck & RSD	155999	0.209131	.564%
11:25	0	Read Baselin	0	0.000000	BL
11:26	1	cal. WH	-628	-0.000841	LOOL
11:26	1	cal. WH	-113	-0.000151	LO
11:27	1	cal. WH	240	0.000321	LO
		!Zweck & RSD	64	0.000085	686%
11:28	0	Read Baselin	0	0.000000	BL
----- Benutzer Nachfrage: Lauf beenden -----					

Seite #1

Ergebnisreport des Laufs
 Ergebnisse: L:\LABOR\CN ANA~1\IMOSER~1\REPORTS\0928_05.RST
 Ergebnisse fertiggestellt: 12:26 September 28, 2015.
 Benutzer: moser

Vi: Halbesee pH 12; WH 9,15; 0,20

Zeit	Prob	Name	Höhe	Kalk.	Flagg
----- Benutzer Nachfrage: Datensammlung beginnen -----					
11:53	0	Carryover	732	0.001040	LO
11:54	0	Baseline	0	0.000000	BL
11:54	1	Cal 0.200 mg	139282	0.197739	LO
11:55	1	Cal 0.200 mg	141298	0.200601	LO
11:56	1	Cal 0.200 mg	141443	0.200808	LO
12:00	1	Cal 0.200 mg	141475	0.200853	LO
		!Zweck & RSD	140874	0.200000	.756%
12:01	0	Blank	-265	-0.000376	LO
12:02	0	Read Baselin	0	0.000000	BL
12:02	5	0,15 pH 12	108206	0.153621	LO
12:03	5	0,15 pH 12	109240	0.155089	LO
12:07	5	0,15 pH 12	108024	0.153362	LO
		!Zweck & RSD	108490	0.154024	.605%
12:08	0	Read Baselin	0	0.000000	BL
12:09	6	0,20 pH 12	144234	0.204770	LO
12:09	6	0,20 pH 12	144971	0.205815	LO
12:10	6	0,20 pH 12	145408	0.206436	LO
		!Zweck & RSD	144871	0.205674	.409%
12:13	0	Read Baselin	0	0.000000	BL
12:16	1	cal. WH	138065	0.196011	LO
12:16	1	cal. WH	139152	0.197555	LO
12:16	1	cal. WH	142006	0.201606	LO
		!Zweck & RSD	139741	0.198391	1.46%
12:17	0	Read Baselin	0	0.000000	BL
12:19	5	0,15 pH 12	107541	0.152676	LO
12:23	5	0,15 pH 12	107650	0.152831	LO
12:23	5	0,15 pH 12	107969	0.153283	LO
		!Zweck & RSD	107720	0.152930	.206%
12:24	0	Read Baselin	0	0.000000	BL
12:25	6	0,20 pH 12	0	0.000000	
12:25	6	0,20 pH 12	0	0.000000	
		!Zweck & RSD	---	---	Kein
----- Benutzer Nachfrage: Lauf beenden -----					

Ergebnisreport des Laufs
 Ergebnisse: L:\LABOR\CN_ANA~1\1MOSER-1\REPORTS\0930_01.RST
 Ergebnisse fertiggestellt: 09:27 September 30, 2015.
 Benutzer: moser

V1

Kanal 2
 Balg p112

Zeit	Prob	Name	Höhe	Kalk.	Flagg
----- Benutzer Nachfrage: Datensammlung beginnen -----					
08:42	0	Carryover	821	0.001233	LO
08:42	0	Baseline	0	0.000000	BL
08:43	1	Cal 0.200 mg	131379	0.197172	LO
08:44	1	Cal 0.200 mg	134066	0.201206	LO
08:44	1	Cal 0.200 mg	134080	0.201226	LO
08:48	1	Cal 0.200 mg	133526	0.200396	LO
		!Zweck & RSD	133263	0.200000	.962%
08:50	0	Blank	-270	-0.000405	LO
08:50	0	Read Baselin	0	0.000000	BL
08:51	6	0,2	136067	0.204208	LO
08:52	6	0,2	137378	0.206175	LO
08:54	6	0,2	137442	0.206271	LO
		!Zweck & RSD	136962	0.205552	.567%
08:57	0	Read Baselin	0	0.000000	BL
08:58	5	0,15	102141	0.153293	LO
08:58	5	0,15	103081	0.154703	LO
08:59	5	0,15	101977	0.153047	LO
		!Zweck & RSD	102400	0.153681	.582%
09:00	0	Read Baselin	0	0.000000	BL
09:01	4	0,10	63828	0.095793	LOOL
09:01	4	0,10	65896	0.098896	LO
09:05	4	0,10	65079	0.097670	LO
		!Zweck & RSD	65487	0.098283	1.59%
09:06	0	Read Baselin	0	0.000000	BL
09:07	3	0,05	31913	0.047895	LO
09:11	3	0,05	32226	0.048364	LO
09:12	3	0,05	33149	0.049750	LOOL
		!Zweck & RSD	32069	0.048130	2%
09:13	0	Read Baselin	0	0.000000	BL
09:13	2	0,02	13029	0.019553	LO
09:14	2	0,02	13243	0.019875	LO
09:18	2	0,02	13365	0.020058	LO
		!Zweck & RSD	13212	0.019829	1.29%
09:19	0	Read Baselin	0	0.000000	BL
09:20	1	cal. WH	131722	0.197688	LO
09:21	1	cal. WH	135903	0.203963	LOOL
09:21	1	cal. WH	132489	0.198839	LO
		!Zweck & RSD	132106	0.198263	1.68%
09:24	0	Read Baselin	0	0.000000	BL
----- Benutzer Nachfrage: Lauf beenden -----					

Ergebnisreport des Laufs
 Ergebnisse: L:\LABOR\CN ANA-1\IMOSER-1\REPORTS\0928_06.RST
 Ergebnisse fertiggestellt: 13:43 September 28, 2015.
 Benutzer: moser

V4: pCO_2 pH7

Zeit	Prob	Name	Höhe	Kalk.	Flagg
----- Benutzer Nachfrage: Datensammlung beginnen -----					
12:55	0	Carryover	765	0.001068	LO
12:59	0	Baseline	0	0.000000	BL
13:00	1	Cal 0.200 mg	141173	0.197128	LO
13:01	1	Cal 0.200 mg	145918	0.203752	LOOL
13:01	1	Cal 0.200 mg	142600	0.199120	LO
		!Zweck & RSD	141887	0.198124	1.72%
13:02	0	Blank	-201	-0.000280	LO
13:06	0	Read Baselin	0	0.000000	BL
13:07	2	0,02 pb ph7	12975	0.018117	LO
13:08	2	0,02 pb ph7	13047	0.018218	LO
13:08	2	0,02 pb ph7	13048	0.018220	LO
		!Zweck & RSD	13023	0.018185	.323%
13:09	0	Read Baselin	0	0.000000	BL
13:13	3	0,05 pb ph7	32436	0.045291	LO
13:14	3	0,05 pb ph7	32588	0.045504	LO
13:15	3	0,05 pb ph7	32667	0.045614	LO
		!Zweck & RSD	32563	0.045470	.361%
13:16	0	Read Baselin	0	0.000000	BL
13:16	4	0,10 pb ph7	66900	0.093415	LO
13:20	4	0,10 pb ph7	66883	0.093393	LO
13:20	4	0,10 pb ph7	66714	0.093156	LO
		!Zweck & RSD	66832	0.093321	.154%
13:22	0	Read Baselin	0	0.000000	BL
13:23	5	0,15 pb ph7	98868	0.138054	LO
13:23	5	0,15 pb ph7	99801	0.139357	LO
13:26	5	0,15 pb ph7	99576	0.139042	LO
		!Zweck & RSD	99415	0.138818	.49%
13:29	0	Read Baselin	0	0.000000	BL
13:30	6	0,20 pb ph7	130982	0.182898	LO
13:30	6	0,20 pb ph7	132520	0.185044	LO
13:31	6	0,20 pb ph7	132634	0.185204	LO
		!Zweck & RSD	132045	0.184382	.699%
13:32	0	Read Baselin	0	0.000000	BL
13:36	1	cal. WH	143686	0.200635	LO
13:37	1	cal. WH	145460	0.203114	LO
13:37	1	cal. WH	145697	0.203444	LO
		!Zweck & RSD	144948	0.202398	.758%
13:38	0	Read Baselin	0	0.000000	BL
----- Benutzer Nachfrage: Lauf beenden -----					

Ergebnisreport des Laufs
 Ergebnisse: L:\LABOR\CN ANA~1\1MOSER~1\REPORTS\0928_07.RST
 Ergebnisse fertiggestellt: 14:39 September 28, 2015.
 Benutzer: moser

12.1 + PCO₂ pH11

Zeit	Prob	Name	Höhe	Kalk.	Flagg
----- Benutzer Nachfrage: Datensammlung beginnen -----					
13:52	0	Carryover	598	0.000783	LO
13:52	0	Baseline	0	0.000000	BL
13:53	1	Cal 0.200 mg	151326	0.198070	LO
13:57	1	Cal 0.200 mg	153489	0.200900	LO
13:58	1	Cal 0.200 mg	153587	0.201029	LO
		!Zweck & RSD	152801	0.200000	.836%
13:59	0	Blank	247	0.000323	LO
14:01	0	Read Baselin	0	0.000000	BL
14:02	2	0,02 pb ph11	13568	0.017759	LO
14:03	2	0,02 pb ph11	13695	0.017925	LO
14:04	2	0,02 pb ph11	13608	0.017812	LO
		!Zweck & RSD	13624	<u>0.017832</u>	.476%
14:04	0	Read Baselin	0	0.000000	BL
14:08	3	0,05 pb ph11	34230	0.044803	LO
14:10	3	0,05 pb ph11	34542	0.045211	LO
14:10	3	0,05 pb ph11	34577	0.045258	LO
		!Zweck & RSD	34450	<u>0.045091</u>	.555%
14:11	0	Read Baselin	0	0.000000	BL
14:11	4	0,10 pb ph11	70661	0.092487	LO
14:15	4	0,10 pb ph11	71321	0.093351	LO
14:15	4	0,10 pb ph11	71078	0.093033	LO
		!Zweck & RSD	71020	<u>0.092857</u>	.47%
14:17	0	Read Baselin	0	0.000000	BL
14:18	5	0,15 pb ph11	103119	0.134971	LO
14:19	5	0,15 pb ph11	105058	0.137510	LO
14:21	5	0,15 pb ph11	105014	0.137452	LO
		!Zweck & RSD	104397	<u>0.136644</u>	1.06%
14:24	0	Read Baselin	0	0.000000	BL
14:25	6	0,20 pb ph11	141770	0.185561	LO
14:25	6	0,20 pb ph11	144395	0.188997	LO
14:26	6	0,20 pb ph11	144233	0.188785	LO
		!Zweck & RSD	143466	<u>0.187781</u>	1.03%
14:27	0	Read Baselin	0	0.000000	BL
14:31	1	cal. WH	152135	0.199129	LO
14:32	1	cal. WH	154328	0.201999	LO
14:32	1	cal. WH	154132	0.201742	LO
		!Zweck & RSD	153532	0.200957	.79%
14:33	0	Read Baselin	0	0.000000	BL
----- Benutzer Nachfrage: Lauf beenden -----					

Ergebnisreport des Laufs
 Ergebnisse: L:\LABOR\CN_ANA~1\1MOSER~1\REPORTS\0930_03.RST
 Ergebnisse fertiggestellt: 11:53 September 30, 2015.
 Benutzer: moser

V2.2 PHU

Kanal 2				
Zeit	Prob	Name	Höhe	Kalk. Flagg
----- Benutzer Nachfrage: Datensammlung beginnen -----				
11:13	0	Carryover	699	0.000884 LO
11:14	0	Baseline	0	0.000000 BL
11:15	1	Cal 0.200 mg	157323	0.198832 LO
11:15	1	Cal 0.200 mg	158570	0.200408 LO
11:20	1	Cal 0.200 mg	158509	0.200330 LO
11:20	1	Cal 0.200 mg	158588	0.200430 LO
		!Zweck & RSD	158247	0.200000 .39%
11:21	0	Blank	-159	-0.000201 LO
11:22	0	Read Baselin	0	0.000000 BL
11:22	2	0,1pB ph11	117160	0.148072 LO
11:26	2	0,1pB ph11	118854	0.150213 LO
11:27	2	0,1pB ph11	118568	0.149852 LO
		!Zweck & RSD	118194	0.149379 .767%
11:28	0	Read Baselin	0	0.000000 BL
11:29	3	0,5pb ph11	130291	0.164667 LO
11:29	3	0,5pb ph11	131566	0.166279 LO
11:33	3	0,5pb ph11	132328	0.167242 LO
		!Zweck & RSD	131395	0.166063 .783%
11:35	0	Read Baselin	0	0.000000 BL
11:35	4	1,0pb ph11	143840	0.181792 LO
11:35	4	1,0pb ph11	144784	0.182984 LO
11:36	4	1,0pb ph11	145519	0.183914 LO
		!Zweck & RSD	144715	0.182897 .582%
11:39	0	Read Baselin	0	0.000000 BL
11:42	5	2,0pb ph11	146301	0.184902 LO
11:42	5	2,0pb ph11	147641	0.186595 LO
11:43	5	2,0pb ph11	146859	0.185607 LO
		!Zweck & RSD	146934	0.185701 .458%
11:44	0	Read Baselin	0	0.000000 BL
11:45	1	cal. WH	156738	0.198092 LO
11:49	1	cal. WH	157832	0.199475 LO
11:50	1	cal. WH	159390	0.201444 LO
		!Zweck & RSD	157987	0.199670 .844%
11:50	0	Read Baselin	0	0.000000 BL
----- Benutzer Nachfrage: Lauf beenden -----				

Ergebnisreport des Laufs
 Ergebnisse: L:\LABOR\CN_ANA~1\1MOSER~1\REPORTS\0930_02.RST
 Ergebnisse fertiggestellt: 10:56 September 30, 2015.
 Benutzer: moser

V.L.L. p17

Kanal 2				
Zeit	Prob	Name	Höhe	Kalk. Flagg
----- Benutzer Nachfrage: Datensammlung beginnen -----				
10:15	0	Carryover	644	0.000820 LO
10:15	0	Baseline	0	0.000000 BL
10:16	1	Cal 0.200 mg	155703	0.198269 LO
10:20	1	Cal 0.200 mg	157126	0.200080 LO
10:21	1	Cal 0.200 mg	157583	0.200662 LO
10:22	1	Cal 0.200 mg	157839	0.200989 LO
		!Zweck & RSD	157063	0.200000 .607%
10:23	0	Blank	164	0.000209 LO
10:23	0	Read Baselin	0	0.000000 BL
10:27	2	0,1pB	117678	0.149848 LO
10:29	2	0,1pB	118837	0.151324 LO
10:29	2	0,1pB	118414	0.150786 LO
		!Zweck & RSD	118310	0.150653 .496%
10:30	0	Read Baselin	0	0.000000 BL
10:30	3	0,5pb	128578	0.163728 LO
10:33	3	0,5pb	130084	0.165646 LO
10:36	3	0,5pb	129479	0.164876 LO
		!Zweck & RSD	129381	0.164750 .586%
10:36	0	Read Baselin	0	0.000000 BL
10:37	4	1,0pb	144238	0.183668 LO
10:37	4	1,0pb	145667	0.185488 LO
10:39	4	1,0pb	144922	0.184540 LO
		!Zweck & RSD	144942	0.184566 .493%
10:43	0	Read Baselin	0	0.000000 BL
10:44	5	2,0pb	153565	0.195546 LOOL
10:44	5	2,0pb	147278	0.187541 LO
10:45	5	2,0pb	147764	0.188159 LO
		!Zweck & RSD	147521	0.187850 2.37%
10:46	0	Read Baselin	0	0.000000 BL
10:50	1	cal. WH	158170	0.201410 LO
10:51	1	cal. WH	159595	0.203225 LO
10:54	1	cal. WH	159664	0.203313 LO
		!Zweck & RSD	159143	0.202649 .53%
10:54	0	Read Baselin	0	0.000000 BL
----- Benutzer Nachfrage: Lauf beenden -----				

11.3. Ad 7.2.4.1: original measurement report "0922_03"

Ergebnisreport des Laufs
 Ergebnisse: L:\LABOR\CN_ANA-1\1MOSER-1\REPORTS\0922_03.RST
 Ergebnisse fertiggestellt: 11:51 September 22, 2015.
 Benutzer: moser

V1 pH7 V1: pH7 vs. pH11

Zeit	Prob	Name	Höhe	Kalk.	Flagg
----- Benutzer Nachfrage: Datensammlung beginnen -----					
11:16	0	Carryover	731	0.001341	LO
11:19	0	Baseline	0	0.000000	BL
11:21	1	Cal 0.200 mg	106851	0.195905	LOOL
11:21	1	Cal 0.200 mg	109915	0.201524	LO
11:22	1	Cal 0.200 mg	110486	0.202571	LO
		!Zweck & RSD	110201	0.202047	0.77%
11:23	0	Blank	158	0.000289	LO
11:25	0	Read Baselin	0	0.000000	BL
11:28	4	0,10 pH 7	49732	0.091181	LO
11:29	4	0,10 pH 7	50320	0.092260	LO
11:29	4	0,10 pH 7	50442	0.092484	LO
		!Zweck & RSD	50165	0.091975	.757%
11:30	0	Read Baselin	0	0.000000	BL
11:32	5	0,15 pH 7	74608	0.136790	LO
11:35	5	0,15 pH 7	75755	0.138893	LO
11:36	5	0,15 pH 7	75220	0.137912	LO
		!Zweck & RSD	75194	0.137865	.763%
11:36	0	Read Baselin	0	0.000000	BL
11:37	6	0,20 pH 7	99553	0.182525	LO
11:38	6	0,20 pH 7	101160	0.185471	LO
11:38	6	0,20 pH 7	100870	0.184940	LO
		!Zweck & RSD	100528	0.184312	.852%
11:43	0	Read Baselin	0	0.000000	BL
11:44	1	cal. WH	108200	0.198380	LO
11:44	1	cal. WH	110252	0.202142	LO
11:45	1	cal. WH	110482	0.202563	LO
		!Zweck & RSD	109645	0.201028	1.15%
11:49	0	Read Baselin	0	0.000000	BL
11:50	6	0,20 pH 7	0	0.000000	
----- Benutzer Nachfrage: Lauf beenden -----					

-> wurde eliminiert?
 -> bezieht sich auf immos nach auf alle?

Ad 7.2.4.1: Validata plot for the comparison of calibration curves at pH 7, pH 11 and pH 12 (V 1)

pH 7:

23.09.2015	SOP	mOSER								Validata 3.02
Beschreibung Konzentrationen aus Messung bei pH11; Signalhöhen aus Messung pH7; OHNE NIEDRIG pH7; c aus pH11; h aus pH7; OHNE NIEDRIGSTEN PUNKT.xls(Komponente1)										
Verfahren Verfahren										
Kalibrationskurve										
NORMGERECHT VALIDIERT										
# Messung	# Rep.	# Konzstufen	Profil		Arbeitsbereich					
12	3	4	voest_01_ohne_Leerprobe		0,043887667	0,181421667				
x	y1	y2	y3	y4	y5	y8	y7	y8	y9	y10
mg/l	pA									
0,043887667	23047	24093	23570							273529
0,088977333	49732	50320	50442							14412,13333
0,135015667	74608	75755	75220							50164,66667
0,181421667	99553	101160	100870							329396,3333
										75194,33333
										733506,3333
										100527,6667

Datenblatt (Linear (normgerecht))										
Konzentration	Meßwerte	geschätzte Werte	Residuen	Vertrauensintervall (-)	Vertrauensintervall (+)	Prognoseintervall (-)	Prognoseintervall (+)	Gewichte	berechnete Konzentration	% Abweichung
0,043887667	23047	24183,27852	-1136,27852	0,042317405	0,045457928	0,041435739	0,046339595		0,041850927	-4,640802211
0,043887667	24093	24183,27852	-90,27852014	0,042317405	0,045457928	0,041435739	0,046339595		0,043725846	-0,368716603
0,043887667	23570	24183,27852	-613,2785201	0,042317405	0,045457928	0,041435739	0,046339595		0,042788386	-2,504759407
0,088977333	49732	49338,39141	393,6085919	0,087942753	0,090011913	0,086828709	0,091125957		0,089682863	0,792931988
0,088977333	50320	49338,39141	981,6085919	0,087942753	0,090011913	0,086828709	0,091125957		0,090738833	1,977469162
0,088977333	50442	49338,39141	1103,608592	0,087942753	0,090011913	0,086828709	0,091125957		0,090555513	2,223240448
0,135015667	74608	75022,75676	-414,756758	0,133986036	0,136045298	0,132869421	0,137161912		0,13427223	-0,550630242
0,135015667	75755	75022,75676	732,243242	0,133986036	0,136045298	0,132869421	0,137161912		0,136328187	0,972124662
0,135015667	75220	75022,75676	197,243242	0,133986036	0,136045298	0,132869421	0,137161912		0,135369218	0,261859733
0,181421667	99553	100912,24	-1359,239981	0,179841717	0,183001616	0,178963523	0,183879811		0,178985276	-1,342943502
0,181421667	101160	100912,24	247,7600195	0,179841717	0,183001616	0,178963523	0,183879811		0,181865768	0,244789524
0,181421667	100870	100912,24	-42,23998051	0,179841717	0,183001616	0,178963523	0,183879811		0,181345953	-0,041733548

Test der Varianzen		
	unten	oben
s(rel)	2,218922359	0,851954364
Freiheitsgrade	2	2
Varianz	273529	733506,3333
Prüfwert	2,681640094	
F_95Var	19	
F_99Var	99	
Ok, kein signifikanter Unterschied auf Niveau 95%		
Ok, kein signifikanter Unterschied auf Niveau 99%		

Linearitätstest	
Prüfwert	7,175302689
F_99	10,56143105
Ok, kein signifikanter Unterschied (99% Niveau)	

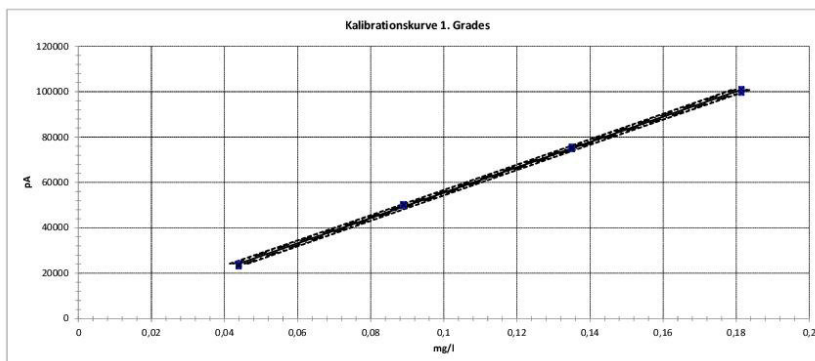
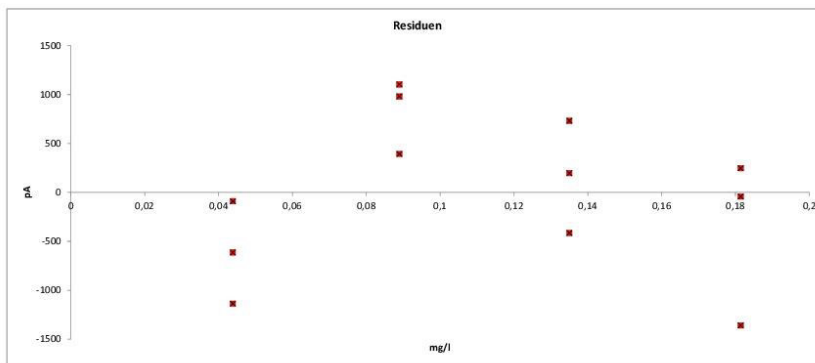
Kalibrationsfunktion 1. Grades (y=a+b*x)		
Steigung	557890,8594	pA/(mg/l)
VB(Steigung)	547646,9404	568134,7784
Achsenabschnitt	-301,2495548	pA
VB(Achsenabschnitt)	-1566,136556	963,6374463
Mittelwert(x)	0,112325583	mg/l
Mittelwert(y)	62364,16667	pA
Reststandardabweichung	816,6783641	pA
Verfahrensstd.abweichung	0,001463868	mg/l
Rel. Verfahrensstd.abweich.	1,30323616	%
s-Wert (95%)	2,228138852	
Qx	0,031553993	(mg/l) ²

Kalibrationsfunktion 2. Grades (y=a+b*x+c*x^2)		
a	-2684,405862	pA
b	611210,9433	pA/mg/l
c	-236516,2282	pA
Empfindlichkeit	558077,2967	pA/(mg/l)
Mittelwert(x)	0,112325583	mg/l
Mittelwert(y)	62364,16667	pA
Reststandardabweichung	842,132782	pA
Verfahrensstd.abweichung	0,001150616	mg/l
Rel. Verfahrensstd.abweich.	1,024358096	%
s-Wert (95%)	2,262157163	
Prüfwert (Lösung)	1,292112063	mg/l
Ok, kein Extremwert innerhalb des Arbeitsbereiches.		

Nachweisgrenze - Kalibrationsmethode		
Ergebnisunsicherheit (k)	3	
Konzentration (l)	0	
Geschätzter Meßwert (l)	-301,2495548	pA
Wiederholungen (Meßprobe)	3	
Entscheidungsniveau NWG	0,95	
t-Werte (1/2-seitig)	1,812461123	2,228138852
Entscheidungsniveau VB	0,95	
Faktoren VB	0,698717043	1,754933527
Kritischer Wert	1036,281055	pA
Nachweisgrenze	0,002397477	mg/l
Schnellschätzung NWG	0,003337618	mg/l
VB Nachweisgrenze	0,001675158	0,004207413 mg/l
Erfassungsgrenze	0,004794954	mg/l
Schnellschätzung EG	0,006675236	mg/l
VB Erfassungsgrenze	0,003350316	0,008414826 mg/l
Bestimmungsgrenze	0,008569394	mg/l
Schnellschätzung BG	0,010012855	mg/l
VB Bestimmungsgrenze	0,005987582	0,015038717 mg/l

Der Arbeitsbereich abgelesen.

ANOVA für Lineare Regression					
Quelle	FG	QS	QS/FG	F-Verhältnis	Wahrsch.
Modell	1	9820934700	9820934700	14724,8447	3,54402E-17
Residuen	10	6669635,504	666963,5504		
LOF	2	3708529,504	1854264,752	5,009654506	0,038851466
FE	8	2961106	370138,25		



Ergebnis Validierung Verfahren

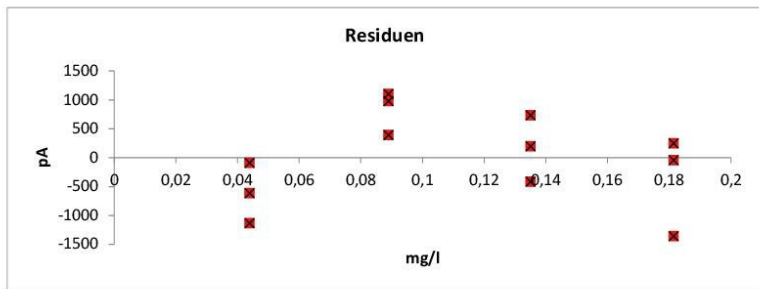
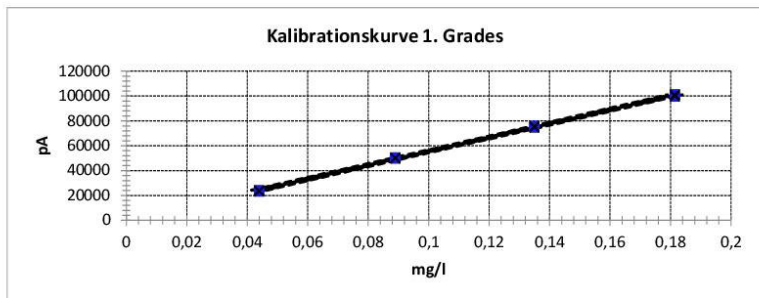
Datum	23.09.2015		
Beschreibung	Konzentrationsdaten aus Messung bei pH11; Signalhöhen aus Messung		
Bearbeiter	mOSER		
# Messung	12	Einheit Konz.:	mg/l
# Rep.	3	Einheit Mess.:	pA
# Konz.stufen	4		

Modell	Linear (normgerecht)
y[pA] =	$557890,85943[\text{pA}/(\text{mg/l})] * x [\text{mg/l}] - 301,24955[\text{pA}]$

Varianzcheck	95% Ok, 99% Ok		
Linearität	Ok		
Varianz unten	273529		
Varianz oben	733506,3333		
VB(Steigung)	547646,9404	568134,7784	pA/(mg/l)
VB(Achsenabschnitt)	-1566,136556	963,6374463	pA
Reststd.abweichung	816,6783641		pA
Verfahrensstd.abweichung	0,001463868		mg/l
Rel. Verfahrensstd.abw.	1,30323616		%

Nachweis- und Bestimmungsgrenze			
	Leerwertmeth.	Kalibriermeth.	
Entscheidungsniveaus (VB)	0,95	0,95	
Nachweisgrenze	N/A	0,002397477	mg/l
Erfassungsgrenze	N/A	0,004794954	mg/l
Bestimmungsgrenze		0,008569394	mg/l

Kalibrierkurve



pH 12:

01.10.2015	SOP	moser		Validata 3.02
Beschreibung		ph12 Kalibrierung		
Verfahren		ANTEC nach ISO Methode		
Kalibrationskurve				
NICHT NORMGERECHT!				
# Messung	# Rep.	# Konstufen	Profil	Arbeitsbereich
15	3	5	voest_01_ohne_Leerprobe	0,02 0,2
x	y1	y2	y3	y4
mg/l	pA			
0,02	15109	15211	15118	
0,05	37760	37973	38159	
0,1	75213	75733	76271	
0,15	114971	119132	116177	
0,2	155024	156737	156235	
				3189 15146
				39861 37964
				279868 75739
				4583397 116760
				775482.3333 155998.6667

Datenblatt (Linear (normgerecht))										
Konzentration	Meßwerte	geschätzte Werte	Residuen	Vertrauensintervall (-)	Vertrauensintervall (+)	Prognoseintervall (-)	Prognoseintervall (+)	Gewichte	berechnete Konzentration	% Abweichung
0,02	15109	14439,23014	669,7698562	0,018564202	0,021435798	0,017561936	0,022438064		0,020853957	4,269786057
0,02	15211	14439,23014	771,7698562	0,018564202	0,021435798	0,017561936	0,022438064		0,020984007	4,920036548
0,02	15118	14439,23014	678,7698562	0,018564202	0,021435798	0,017561936	0,022438064		0,020865432	4,327161101
0,05	37760	37968,62414	-208,6241401	0,048856551	0,051143449	0,047721797	0,052278203		0,049734004	-0,531991959
0,05	37973	37968,62414	4,375859912	0,048856551	0,051143449	0,047721797	0,052278203		0,050005579	0,011158451
0,05	38159	37968,62414	190,3758599	0,048856551	0,051143449	0,047721797	0,052278203		0,050242729	0,485458809
0,1	75213	77184,2808	-1971,280801	0,09911714	0,10088286	0,097840814	0,102159186		0,097486615	-2,513384919
0,1	75733	77184,2808	-1451,280801	0,09911714	0,10088286	0,097840814	0,102159186		0,098149616	-1,850384418
0,1	76271	77184,2808	-913,2808005	0,09911714	0,10088286	0,097840814	0,102159186		0,098835566	-1,1644339
0,15	114971	116399,9375	-1428,937461	0,148922095	0,151077905	0,147753999	0,152246001		0,148178103	-1,214597759
0,15	119132	116399,9375	2732,062539	0,148922095	0,151077905	0,147753999	0,152246001		0,153483382	2,322254912
0,15	116177	116399,9375	-222,9374609	0,148922095	0,151077905	0,147753999	0,152246001		0,149715755	-0,189496985
0,2	155024	155615,5941	-591,5941213	0,198433195	0,201566805	0,197482557	0,202517443		0,199245717	-0,377141537
0,2	156737	155615,5941	1121,405879	0,198433195	0,201566805	0,197482557	0,202517443		0,201429794	0,714896788
0,2	156235	155615,5941	619,4058787	0,198433195	0,201566805	0,197482557	0,202517443		0,200789743	0,394871546

Test der Varianzen		
	unten	oben
s(re)	0,372845845	0,564501458
Freiheitsgrade	2	2
Varianz	3189	775482,3333
Prüfwert	243,1741403	
F_95Var	19	
F_99Var	99	
WARNUNG: Signifikanter Unterschied auf Niveau 95%		
WARNUNG: Signifikanter Unterschied auf Niveau 99%		

Linearitätstest	
Prüfwert	3,71915874
F_99	9,330212103
OK, kein signifikanter Unterschied (99% Niveau)	

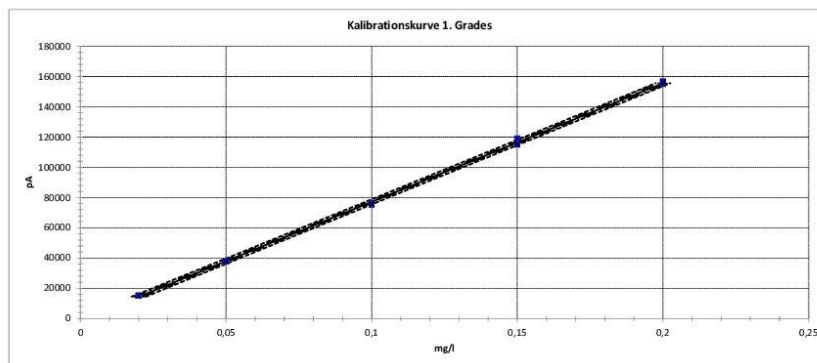
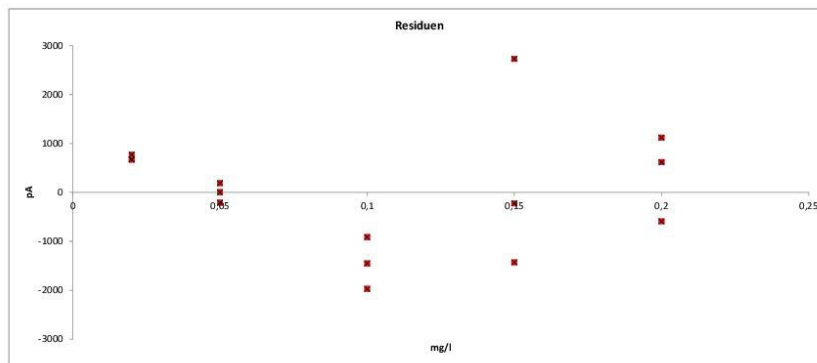
Kalibrierfunktion 1. Grades (y=a+b*x)		
Steigung	784313,1332	pA/(mg/l)
VB(Steigung)	773728,8982	pA/(mg/l)
Achsenabschnitt	-1247,03252	pA
VB(Achsenabschnitt)	-2546,783478	pA
Mittelwert(x)	0,104	mg/l
Mittelwert(y)	80321,53333	pA
Reststandardabweichung	1239,04145	pA
Verfahrensstd.abweichung	0,001579779	mg/l
Rel. Verfahrensstd.abweich.	1,51901831	%
t-Wert (95%)	2,160368556	
Ox	0,06396	(mg/l)*2

Kalibrierfunktion 2. Grades (y=a+b*x+c*x^2)		
a	61,36418633	pA
b	747039,0038	pA/mg/l
c	170299,2539	pA
Empfindlichkeit	782461,2486	pA/(mg/l)
Mittelwert(x)	0,104	mg/l
Mittelwert(y)	80321,53333	pA
Reststandardabweichung	1126,789682	pA
Verfahrensstd.abweichung	0,001440058	mg/l
Rel. Verfahrensstd.abweich.	1,384671278	%
t-Wert (95%)	2,17881283	
Prüfwert (Lösung)	-2,193312615	mg/l
OK, kein Extremwert innerhalb des Arbeitsbereiches		

Nachweisgrenze - Kalibrationsmethode		
Ergebnisunsicherheit (k)	3	
Konzentration (l)	0	
Geschätzter Meßwert (l)	-1247,03252	pA
Wiederholungen (Meßprobe)	3	
Entscheidungsniveau NWG	0,95	
t-Werte (1/2-seitig)	1,770933396	2,160368656
Entscheidungsniveau VB	0,95	
Faktoren VB	0,724953932	1,611042417
Kritischer Wert	408,2972567	pA
Nachweisgrenze	0,002110547	mg/l
Schnellschätzung NWG	0,003412323	mg/l
VB Nachweisgrenze	0,001530049	0,003400181 mg/l
Erfassungsgrenze	0,004221094	mg/l
Schnellschätzung EG	0,006824645	mg/l
VB Erfassungsgrenze	0,003060099	0,006800362 mg/l
Bestimmungsgrenze	0,007587308	mg/l
Schnellschätzung BG	0,010236968	mg/l
VB Bestimmungsgrenze	0,005500449	0,012223475 mg/l

Der Arbeitsbereich abgelesen.

ANOVA für Lineare Regression					
Quelle	FG	QS	QS/FG	F-Verhältnis	Wahrsch.
Modell	1	39344807935	39344807935	25628,06159	8,30363E-23
Residuen	13	19957908,3	1535223,715		
LOF	3	8594313,635	2864771,212	2,521007917	0,117113856
FE	10	11363594,67	1136359,467		



Ergebnis Validierung ANTEC nach ISO Methode

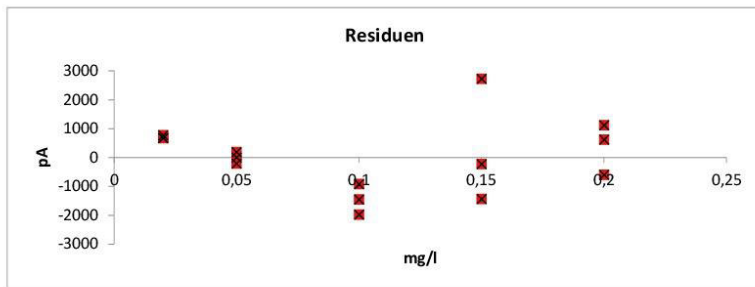
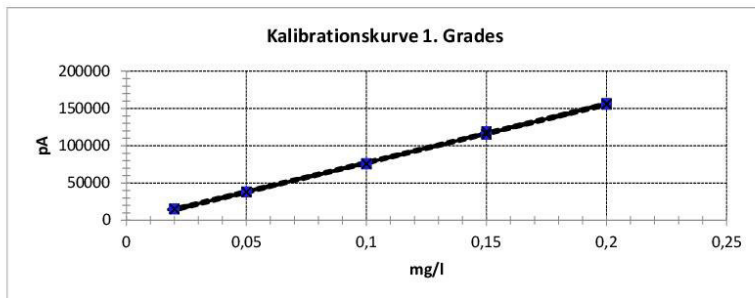
Datum	01.10.2015		
Beschreibung	ph12 Kalibrierung		
Bearbeiter	moser		
# Messung	15	Einheit Konz.:	mg/l
# Rep.	3	Einheit Mess.:	pA
# Konz.stufen	5		

Modell	Linear (normgerecht)
y[pA] = 784313,13321[pA/(mg/l)] * x [mg/l] -1247,03252[pA]	

Varianzcheck	95% nicht ok! 99% nicht ok!		
Linearität	Ok		
Varianz unten	3189		
Varianz oben	775482,3333		
VB(Steigung)	773728,8982	794897,3682	pA/(mg/l)
VB(Achsenabschnitt)	-2546,783478	52,71843779	pA
Reststd.abweichung	1239,04145		pA
Verfahrensstd.abweichung	0,001579779		mg/l
Rel. Verfahrensstd.abw.	1,51901831		%

Nachweis- und Bestimmungsgrenze			
	Leerwertmeth.	Kalibriermeth.	
Entscheidungsniveaus (VB)	0,95	0,95	
Nachweisgrenze	N/A	0,002110547	mg/l
Erfassungsgrenze	N/A	0,004221094	mg/l
Bestimmungsgrenze		0,007587308	mg/l

Kalibrierkurve



Ad 7.2.4.2. Validata report for 0.02 – 0.2 mg/L Cyanide + 0.2 g PbCO3/100mL (V 2.1)

01.10.2015		SOP		moser				Validata 3.02			
Beschreibung		PbCO3 Behandlung; Kalibrationsgerade				V2.1 PbCO3 Behandlung.xls[Komponente1]					
Verfahren		Verfahren									
Kalibrationskurve											
NICHT NORMGERECHT!											
# Messung	# Rep.	# Konzstufen	Profil		Arbeitsbereich						
1S	3	5	voest_01_ohne_Leerprobe		0.02	0.2					
x	y1	y2	y3	y4	y5	y6	y7	y8	y9	y10	
mg/l	µA										
0.02	14552,48991	14633,24361	14634,36519							2204,329956	14606,69957
0.05	35967,2662	36135,81424	36223,41487							16949,07656	36108,33177
0.1	71968,75975	71950,47172	71768,66723							12237,38891	71895,96624
0.15	108531,6589	109555,8532	109308,861							285670,346	109132,1244
0.2	139505,0617	141143,1401	141264,5582							965645,1478	140637,5867

Datenblatt (Linear (normgerecht))										
Konzentration	Meßwerte	geschätzte Werte	Residuen	Vertrauensintervall (-)	Vertrauensintervall (+)	Prognoseintervall (-)	Prognoseintervall (+)	Gewichte	berechnete Konzentration	% Abweichung
0.02	14552,48991	15152,76846	-600,2785895	0,018181624	0,021818376	0,016912297	0,023087703		0,019150026	-4,249869077
0.02	14633,24361	15152,76846	-519,5248572	0,018181624	0,021818376	0,016912297	0,023087703		0,019264371	-3,678146745
0.02	14634,36519	15152,76846	-518,4032781	0,018181624	0,021818376	0,016912297	0,023087703		0,019265959	-3,670206157
0.05	35967,2662	36339,7232	-372,4570046	0,048551819	0,051448181	0,047114754	0,052885246		0,049472614	-1,054772644
0.05	36135,81424	36339,7232	-203,9089573	0,048551819	0,051448181	0,047114754	0,052885246		0,049711272	-0,577456156
0.05	36223,41487	36339,7232	-116,3083274	0,048551819	0,051448181	0,047114754	0,052885246		0,049835311	-0,329377191
0.1	71968,75975	71651,31443	317,445324	0,098881896	0,101118104	0,097265484	0,102734516		0,100449492	0,449491672
0.1	71950,47172	71651,31443	299,1572983	0,098881896	0,101118104	0,097265484	0,102734516		0,100423596	0,423596456
0.1	71768,66723	71651,31443	117,3528081	0,098881896	0,101118104	0,097265484	0,102734516		0,100166168	0,166167544
0.15	108531,6589	106962,9057	1568,753259	0,14863488	0,15136512	0,147155536	0,152844464		0,152221301	1,480867145
0.15	109555,8532	106962,9057	2592,947515	0,14863488	0,15136512	0,147155536	0,152844464		0,153671525	2,447683065
0.15	109308,861	106962,9057	2345,955331	0,14863488	0,15136512	0,147155536	0,152844464		0,153321792	2,2145281
0.2	139505,0617	142274,4969	-2789,435129	0,198015709	0,201984291	0,196811767	0,203188233		0,196078575	-1,960712497
0.2	141143,1401	142274,4969	-1131,356729	0,198015709	0,201984291	0,196811767	0,203188233		0,198398038	-0,800981113
0.2	141264,5582	142274,4969	-1009,936894	0,198015709	0,201984291	0,196811767	0,203188233		0,198569962	-0,715019247

Test der Varianzen		
	unten	oben
f(nel)	0,321429949	0,693726759
Freiheitsgrade	2	2
Varianz	2204,329956	965645,1478
Prüfwert	438,067425	
F_95Var	19	
F_99Var	99	
WARNUNG: Signifikanter Unterschied auf Niveau 95%		
WARNUNG: Signifikanter Unterschied auf Niveau 99%		

Linearitätstest	
Prüfwert	11,09244484
F_99	9,330212103
WARNUNG: Signifikanter Unterschied (99% Niveau)	

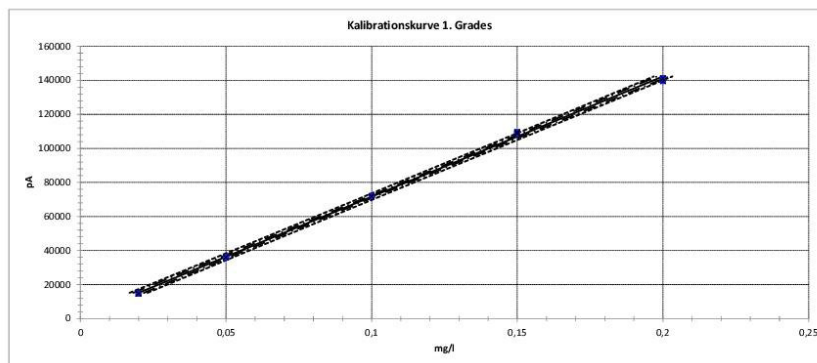
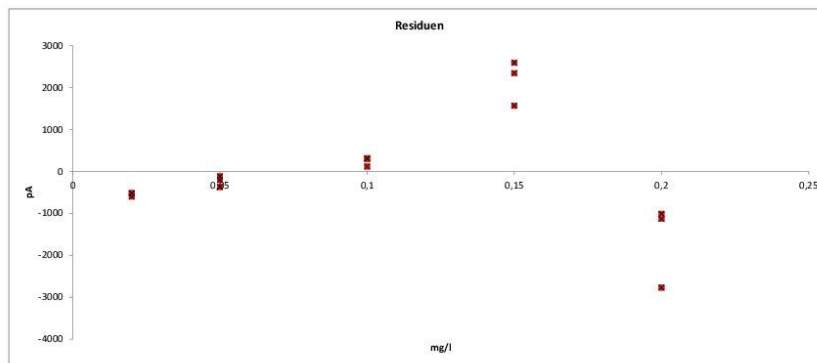
Kalibrierfunktion 1. Grades (y=ax+b)		
Steigung	706,231,8245	µA/(mg/l)
VB(Steigung)	694161,8107	µA/(mg/l)
Achsenabschnitt	1028,131974	µA
VB(Achsenabschnitt)	-454,0735841	2510,337532
Mittelwert(x)	0,104	mg/l
Mittelwert(y)	74476,24172	µA
Reststandardabweichung	1412,973857	µA
Verfahrensstd.abweichung	0,002000722	mg/l
Rel. Verfahrensstd.abweich.	1,923771574	%
t-Wert (95%)	2,160368656	
Qx	0,06396	(mg/l)*2

Kalibrierfunktion 2. Grades (y=ax²+bx+c)		
a	-1097,845738	µA
b	766797,529	µA/mg/l
c	-276714,5594	µA
Empfindlichkeit	709240,9007	µA/(mg/l)
Mittelwert(x)	0,104	mg/l
Mittelwert(y)	74476,24172	µA
Reststandardabweichung	1060,159611	µA
Verfahrensstd.abweichung	0,001494779	mg/l
Rel. Verfahrensstd.abweich.	1,43728778	%
t-Wert (95%)	2,17881283	
Prüfwert (Lösung)	1,385538822	mg/l
OK, kein Extremwert innerhalb des Arbeitsbereiches		

Nachweisgrenze - Kalibrationsmethode		
Ergebnisunsicherheit (k)	3	
Konzentration (l)	0	
Geschätzter Meßwert (l)	1028,131974	pA
Wiederholungen (Meßprobe)	3	
Entscheidungsniveau NWG	0,95	
t-Werte (1/2-seitig)	1,770933396	2,160368656
Entscheidungsniveau VB	0,95	
Faktoren VB	0,724953932	1,611042417
Kritischer Wert	2915,831291	pA
Nachweisgrenze	0,002672917	mg/l
Schnellschätzung NWG	0,00432156	mg/l
VB Nachweisgrenze	0,001937742	0,004306183 mg/l
Erfassungsgrenze	0,005345835	mg/l
Schnellschätzung EG	0,008643121	mg/l
VB Erfassungsgrenze	0,003875484	0,008612367 mg/l
Bestimmungsgrenze	0,009564205	mg/l
Schnellschätzung BG	0,012964681	mg/l
VB Bestimmungsgrenze	0,006933608	0,015408341 mg/l

Der Arbeitsbereich abgelesen.

ANOVA für Lineare Regression					
Quelle	FG	QS	QS/FG	F-Verhältnis	Wahrsch.
Modell	1	31900806421	31900906421	15978,45449	1,78705E-21
Residuen	13	25954436,56	1996495,12		
LOF	3	23389023,98	7796341,328	30,39020463	2,44649E-05
FE	10	2565412,58	256541,258		



Ergebnis Validierung Verfahren

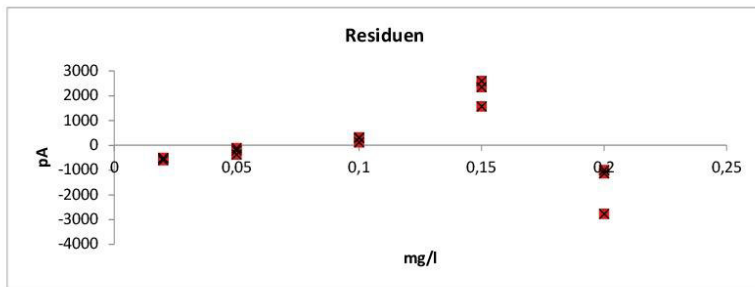
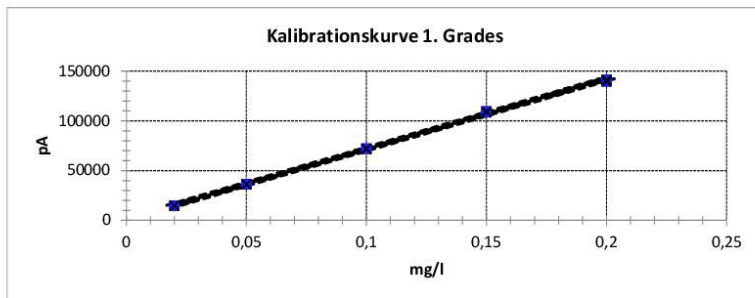
Datum	01.10.2015		
Beschreibung	PbCO3 Behandlung; Kalibrationsgerade		
Bearbeiter	moser		
# Messung	15	Einheit Konz.:	mg/l
# Rep.	3	Einheit Mess.:	pA
# Konz.stufen	5		

Modell	Linear (normgerecht)
y[pA] = 706231,82451[pA/(mg/l)] * x [mg/l] + 1028,13197[pA]	

Varianzcheck	95% nicht ok! 99% nicht ok!		
Linearität	nicht ok!		
Varianz unten	2204,329956		
Varianz oben	965645,1478		
VB(Steigung)	694161,8107	718301,8383	pA/(mg/l)
VB(Achsenabschnitt)	-454,0735841	2510,337532	pA
Reststd.abweichung	1412,973857 pA		
Verfahrensstd.abweichung	0,002000722 mg/l		
Rel. Verfahrensstd.abw.	1,923771574 %		

Nachweis- und Bestimmungsgrenze			
	Leerwertmeth.	Kalibriermeth.	
Entscheidungsniveaus (VB)	0,95	0,95	
Nachweisgrenze	N/A	0,002672917	mg/l
Erfassungsgrenze	N/A	0,005345835	mg/l
Bestimmungsgrenze		0,009564205	mg/l

Kalibrierkurve



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